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FUTURE GLOBAL MORTALITY FROM CHANGES IN AIR POLLUTION ATTRIBUTABLE TO CLIMATE CHANGE

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Ground-level ozone and fine particulate matter (PM_{2.5}) are associated with premature human mortality^{1–4}; their future concentrations depend on changes in emissions, which dominate the near-term⁵, and on climate change^{6,7}. Previous global studies of the air quality-related health effects of future climate change^{8,9} used single atmospheric models. However, in related studies, mortality results differ among models^{10–12}. Here we use an ensemble of global chemistry-climate models¹³ to show that premature mortality from changes in air pollution attributable to climate change, under the high greenhouse gas scenario RCP8.5¹⁴,

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Author contributions: JJW, JFL, DTS and RAS conceived the study. All other co-authors conducted the model simulations. RAS processed model output and estimated human mortality. RAS and JJW analyzed results. RAS and JJW prepared the manuscript and all co-authors commented on it.

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is likely positive. We estimate 3,340 (−30,300 to 47,100) ozone-related deaths in 2030, relative to 2000 climate, and 43,600 (−195,000 to 237,000) in 2100 (14% of the increase in global ozone-related mortality). For PM_{2.5}, we estimate 55,600 (−34,300 to 164,000) deaths in 2030 and 215,000 (−76,100 to 595,000) in 2100 (countering by 16% the global decrease in PM_{2.5}-related mortality). Premature mortality attributable to climate change is estimated to be positive in all regions except Africa, and is greatest in India and East Asia. Most individual models yield increased mortality from climate change, but some yield decreases, suggesting caution in interpreting results from a single model. Climate change mitigation will likely reduce air pollution-related mortality.

Climate change can affect air quality through several pathways, including changes in the ventilation and dilution of air pollutants, photochemical reaction rates, removal processes, stratosphere–troposphere exchange of ozone, wildfires, and natural biogenic and lightning emissions^{6,7}. Overall, changes in these processes are expected to increase ozone in polluted regions during the warm season, especially in urban areas and during pollution episodes, but decrease ozone in remote regions due to greater water vapour concentrations leading to greater ozone destruction. These effects are exacerbated by the greater decomposition of reservoir species such as PAN⁷. PM_{2.5} will also be affected by climate change, but impacts vary in sign among models and show regional variation related to differences in precipitation, wildfires, biogenic emissions, PM_{2.5} composition, and other factors.

Previous studies have examined the impact of future climate change on human health via air quality globally^{8–9,15}, in the US^{10, 16–20}, and in Europe²¹. However, only two studies have previously used an ensemble of models to assess air pollution-related mortality attributable to climate change: one for the US¹⁰, and our previous global work with the same ensemble used here, but evaluating the effects of historical climate change prior to 2000¹¹. Both studies found a large spread of mortality outcomes depending on the atmospheric model used. Silva et al.¹¹ found that the multi-model average suggested a small detrimental effect of climate change on global present-day air pollution-related mortality, but individual models yielded estimates of opposing sign.

The Atmospheric Chemistry and Climate Model Intercomparison Project (ACCMIP) ensemble (Supplementary Table 1) simulated air quality in 2000, and in 2030, 2050 and 2100 for the four global Representative Concentration Pathway scenarios (RCPs)²². We previously estimated future air pollution premature mortality under all four RCP scenarios, estimating the net effect of both emissions changes and climate change¹². Under RCP8.5, ozone concentrations increase in most locations in 2100 relative to 2000, due to increases in methane emissions and the effect of climate change^{7,23}, but PM_{2.5} decreases in 2100 due to a projected decrease in particulate and precursor emissions²⁴. These changes in pollutant concentrations lead to 316,000 (95% C.I.: −187,000 to 1.38 million) ozone-related excess deaths yr^{−1} and −1.31 (−2.04 to −0.17) million PM_{2.5}-related (avoided) deaths yr^{−1} in 2100¹². Here we present results from additional ACCMIP simulations that were designed to isolate the influences of future climate change under RCP8.5, by simulating the projected climates of 2030 and 2100 (imposed by prescribing sea-surface temperatures, sea ice cover, and greenhouse gas concentrations for radiation) together with air pollutant emissions from 2000. The effects of climate change are then isolated by a difference with historical 2000

simulations. Premature mortality attributable to RCP8.5 climate change is estimated following the methods of Silva et al.¹², including projected population and baseline mortality rates (see Methods), such that mortality estimates here can be compared directly with overall changes in air pollution-related mortality in RCP8.5.

We estimate that global ozone mortality attributable to RCP8.5 climate change will be 3,340 (−30,300 to 47,100) deaths yr^{−1} in 2030 and 43,600 (−195,000 to 237,000) deaths yr^{−1} in 2100 (Figures 1a and 2a). In 2100, ozone mortality increases in most regions, especially in highly populated and highly polluted areas, with marked spatial differences within regions that include both positive and negative mortality changes (Figure 3a, Supplementary Table 2, Supplementary Figures 1 and 2a). The effect on ozone mortality in 2100 is greatest in East Asia (45,600 deaths yr^{−1}, 41 deaths yr^{−1} per million people), India (16,000 deaths yr^{−1}, 8 deaths yr^{−1} per million people) and North America (9,830 deaths yr^{−1}, 13 deaths yr^{−1} per million people), but some areas within these and other regions show decreases in mortality. East Asia has high mortality effects per person in part because of its higher projected mortality rate from respiratory diseases. Climate change contributes 14% of the overall increase in ozone mortality estimated for RCP8.5 in 2100 relative to 2000¹². However, three of 8 models in 2030 and three of 9 in 2100 show global decreases in ozone mortality due to climate change. For each model, the uncertainty range does not include zero; only the spread of models causes the overall uncertainty to span zero. Uncertainty in modeled ozone concentrations contributes over 97% to the overall uncertainty in both 2030 and 2100, with the remainder from uncertainties in relative risk (RR). Results from a sensitivity analysis using present-day population and baseline mortality rates (Table 1) show 32% and 67% lower mortality estimates in 2030 and 2100, respectively, largely because the projected baseline mortality rates of chronic respiratory diseases increase through 2100. The models agree that ozone will increase due to climate change in some polluted regions, notably the northeast US as found in other studies⁶ and decrease in the tropics over the oceans (Supplementary Figures 3 and 4a). These changes are consistent with those analysed by Schnell et al.²⁵ for 2100, using four of these same models, and were attributed to a greater efficiency of precursor emissions to generate surface ozone in polluted regions, along with reductions in the export of precursors to downwind regions.

The impact of climate change on PM_{2.5} mortality is estimated to result in 55,600 (−34,300 to 164,000) deaths yr^{−1} in 2030 and 215,000 (−76,100 to 595,000) deaths yr^{−1} in 2100 (Figures 1b and 2b). Mean estimates of PM_{2.5} mortality increase in 2100 in all regions except Africa (−25,200 deaths yr^{−1}) (Figure 3b, Supplementary Table 3, Supplementary Figure 2b). The greatest increases in mortality in 2100 occur in India (80,200 deaths yr^{−1}, 40 deaths yr^{−1} per million people), Middle East (50,400 deaths yr^{−1}, 45 deaths yr^{−1} per million people) and East Asia (47,200 deaths yr^{−1}, 43 deaths yr^{−1} per million people), although the Former Soviet Union shows greater mortality per million people in 2100 (11,800 deaths yr^{−1}, 57 deaths yr^{−1} per million people). Similar to ozone mortality, there are substantial spatial differences within each region, including both increases and decreases in mortality. For PM_{2.5}, a large decrease in mortality is projected in RCP8.5 relative to 2000 (when accounting for changes in both emissions and climate)¹², but climate change alone increases mortality, partially counteracting the decrease associated with declining emissions in RCP8.5. Without climate change, the decrease in PM_{2.5}-related mortality would be roughly

16% greater in 2100 relative to 2000. Propagating uncertainty in RR to the mortality estimates leads to coefficients of variation (CVs) of 8–31% (2030) and 11–46% (2100) for the different models, but the spread of model results increases overall CVs to 123% in 2030 and 106% in 2100. In both years, one model (GISS-E2-R) yields a decrease in global mortality from climate change while the other three (2030) or four (2100) show an increase. Uncertainty in modeled PM_{2.5} concentrations in 2000 makes a similar contribution to the overall uncertainty (50% in 2030 and 52% in 2100) compared with uncertainty in modeled PM_{2.5} concentrations in future years (50% in 2030, 48% in 2100). Uncertainty in RR makes a negligible contribution in both periods (<1%), as the multi-model mean is small and different models disagree on the sign of the influence. Considering present-day population and baseline mortality rates (Table 1), we estimate 23% and 33% lower mortality in 2030 and 2100, respectively, mostly associated with the increase in projected baseline mortality rates through 2100.

PM_{2.5}-related mortality was estimated above for the sum of PM_{2.5} species reported by five models, using a common formula (see Methods), to increase the number of models considered and to increase consistency among PM_{2.5} estimates. Additionally, we present a sensitivity analysis considering the PM_{2.5} concentrations reported by four models using their own PM_{2.5} formulas, for which multi-model average mortality results are modestly higher: 15% greater in 2030 and 12% in 2100 (Supplementary Figure 5). The degree of agreement between the two estimates varies among the four models, and for one model (GISS-E2-R) the two sources of PM_{2.5} estimates yield impacts of different sign in 2030.

There is considerable agreement among models regarding the increase in PM_{2.5} concentrations in many locations in 2100, including most polluted regions, due to RCP8.5 climate change (Supplementary Figure 4b). Allen et al.²⁶ analysed four of these same models in 2100 and found that global average surface PM_{2.5} concentrations increased due to climate change, reflecting increases in nearly all relevant species for each model. They attributed this increase in PM_{2.5} mainly to a decrease in wet deposition associated with less large-scale precipitation over land. Our multi-model mean estimates of global population-weighted changes for PM_{2.5} and individual species (Supplementary Table 4; Supplementary Figure 6) are similar to those of Allen et al.²⁶. Unlike Allen et al.²⁶, however, GISS-E2-R shows a net decrease in global population-weighted concentrations of total PM_{2.5} and of each PM_{2.5} species except sea salt, in 2100, likely due to projected concentration decreases over densely-populated eastern China. Models also differ strongly in the sign and magnitude of changes in dust, particularly over North Africa and the Middle East; HadGEM2 projects increases in PM_{2.5} for all species except dust, but a strong decrease in dust over the Middle East and South Asia. In Africa, the decrease in PM_{2.5} near the equator is likely caused by increased precipitation, whereas PM_{2.5} increases are associated with precipitation decreases in Southern Africa²⁶. Differences in PM_{2.5} (and ozone) responses to climate change among models likely result from differences in large-scale meteorological changes, and different treatments of atmospheric chemistry and feedback processes among the models (such as the response of dust to climate change).

In the US, our multi-model mean mortality estimates for the impact of RCP8.5 climate change for ozone (1,130 deaths yr⁻¹ in 2030; 8,810 deaths yr⁻¹ in 2100) compare well with

those of Fann et al.²⁰, who report 420 to 1900 ozone-related deaths yr⁻¹ for RCP8.5 climate change in 2030, despite differences in concentration-response functions and population and baseline mortality projections. These results for ozone and those for PM_{2.5} (6,900 deaths yr⁻¹ in 2030; 19,400 deaths yr⁻¹ in 2100) are also consistent with the increases in mortality and spatial heterogeneity attributed to climate change in 2050 by Bell et al.¹⁶ for ozone and Tagaris et al.¹⁷ for ozone and PM_{2.5}, although these studies used different climate change scenarios besides other methodological differences. Across models, our estimates for ozone mortality in the US vary between -435 and 4,750 deaths yr⁻¹ in 2030 and between -1,820 and 27,012 deaths yr⁻¹ in 2100. This spread of model results, with a few models suggesting avoided mortality due to climate change, is similar to that of Post et al.¹⁰ (-600 to 2,500 deaths yr⁻¹ in 2050) using SRES scenarios of GHG emissions. Similarly, results show spatial heterogeneity within several regions (Figure 2) that is similar to Post et al.¹⁰ for the US and Orru et al.²¹ for Europe.

The spread of results among models highlights the uncertainty in the effect of climate change on air quality. Further improvements in chemistry climate models are needed to better model the interaction and feedbacks between climate and air quality, including the sensitivity of biogenic emissions to climate change, the effects of meteorological changes on air quality (e.g., aerosol-cloud interactions, secondary aerosol formation, wet deposition, and gas-aerosol partitioning), and the impact of climate change on wildfires. Stratosphere-troposphere exchange of ozone is also important, as is the impact of land use changes on regional climate and air pollution. Our results are specific to climate change as projected under RCP8.5 and would differ for other scenarios. We estimate the effect of climate change as the difference between simulations with future climate and year 2000 climate, both with year 2000 emissions, although global emissions of PM_{2.5} and its main precursors decrease under RCP8.5. Had we instead modelled future emissions with present vs. future climate, we would likely have attributed smaller changes in air pollution and mortality to climate change, given the projected emission reductions. Whereas the net effect of missing and uncertain processes does not clearly indicate an under- or overestimate for the effect of climate change on air quality, we likely underestimate the magnitude of the health impact by omitting mortality for people under 25, and morbidity effects. We also neglect possible synergistic effects of a warmer climate to modify air pollution-mortality relationships. Although a few studies have suggested stronger relationships between ozone²⁷ and PM_{2.5}²⁸ and health at higher temperatures, there is insufficient evidence to include those effects here.

Despite these uncertainties, this study is the first to use a multi-model ensemble to show that global air pollution-related mortality attributable to climate change is likely positive. The spread of results among models within the ensemble, including differences in the sign of global and regional mortality estimates, suggests that results from studies using a single model and a small number of model years should be interpreted cautiously. Actions to mitigate climate change, such as reductions in long-lived GHG emissions, will likely benefit human health by reducing the effect of climate change on air quality in many locations. These health benefits are likely to be smaller than those from reducing co-emitted air pollutants²⁹, but both types of health benefits via changes in air quality would add to reductions in many other influences of climate change on human health³⁰.

Methods

The Atmospheric Chemistry and Climate Model Intercomparison Project (ACCMIP)¹³ included contributions from 14 modelling groups, of which 9 completed simulations that are used here (Supplementary Table 1). ACCMIP models incorporate chemistry-climate interactions, including mechanisms by which climate change affects ozone and PM_{2.5}, although models do not all include the same interactions, and do not always agree on their net effects⁷. Of these nine, three models are not truly coupled chemistry-climate models: MOCAGE is a chemical transport model driven by external meteorology, and UM-CAM and STOC-HadAM3 do not model the feedback of chemistry on climate¹³. As a result, these models do not fully capture the effects of changes in air pollutant concentrations on processes that affect meteorology, such as through radiative transfer and clouds. Prescribed anthropogenic and biomass burning emissions were very similar for the different models, but they used different natural emissions (e.g. biogenic volatile organic compounds, ocean emissions, soil and lightning NO_x)^{14, 23}. Modelled 2000 concentrations show good agreement with observations for ozone²³ and PM_{2.5}²⁴, although models tend to overestimate ozone in the Northern Hemisphere and underestimate it in the Southern Hemisphere, and to underestimate PM_{2.5}, particularly in East Asia.

We isolate the effect of climate change on air quality as the difference in concentrations between ACCMIP simulations using year 2000 emissions together with future year climate, imposed by prescribing RCP8.5³¹ sea surface temperatures, sea ice cover, and GHGs (for radiation) for 2030 and 2100 (referred to as “Em2000CI2030” and “Em2000CI2100”), and simulations with 2000 emissions and climate (“acchist2000”)¹³. We analyse results from the nine models reporting ozone from the Em2000CI2030/2100 simulations, and the five reporting PM_{2.5} (Supplementary Table 1). Ozone and PM_{2.5} species surface concentrations from each model are calculated in each grid cell, after regridding output from the native horizontal resolutions of each model (1.9°×1.2° to 5°×5°) to a common 0.5°×0.5° resolution. To be consistent with the epidemiological studies considered^{1,4}, we use the seasonal average of daily 1-hr maximum ozone concentrations for the six consecutive months with highest concentrations in each grid cell, and annual average PM_{2.5} concentration.

Seven of the nine models with Em2000CI2030/2100 simulations reported both hourly and monthly ozone concentrations, while two reported only monthly values. We calculate the ratio of the 6-month average of daily 1-hr maximum concentrations to the annual average concentrations, for each grid cell and each year, for those models that reported both hourly and monthly concentrations; then, we apply that ratio to the annual average ozone concentrations for the other two models, following Silva *et al.*^{11,12}.

We calculate PM_{2.5} concentration using the sum of PM_{2.5} species mass mixing ratios reported by five models and a common formula:

$$\text{PM}_{2.5} = \text{BC} + \text{OA} + \text{SO}_4 + \text{SOA} + \text{NH}_4 + 0.25 * \text{SS} + 0.1 * \text{Dust},$$

where BC – Black Carbon, OA – (Primary) Organic Aerosol corrected to include species other than carbon, NH₄ – NH₄ in ammonium sulfate, SOA – Secondary Organic Aerosol,

and SS – Sea Salt, as had been done previously by Fiore et al.³³ and Silva *et al.*^{11,12}. The factors 0.25 and 0.1 are intended to approximate the fractions of sea salt and dust that are in the PM_{2.5} size range. Nitrate was reported by three models, but we chose to omit nitrate from our PM_{2.5} formula to avoid imposing changes inconsistent with the effect of climate change for other models, following Silva *et al.*¹¹, although nitrate was included in estimates of total PM_{2.5} by Silva *et al.*¹². Four of these models also reported their own estimate of PM_{2.5} (Supplementary Table 1).

The impact of climate change on global population-weighted differences (Em2000CI2030/2100 minus acchist2000) in PM_{2.5} and ozone concentrations for the different models are shown in Supplementary Tables 4 and 5, respectively, while regional multi-model average differences are shown in Supplementary Figures 7 and 8.

We estimate premature mortality by calculating the fraction of cause-specific mortality attributable to long-term changes in pollutant concentrations, using methods that are identical to those of Silva *et al.*¹², so that mortality attributable to climate change can be compared simply with changes in mortality under the RCP scenarios. We use relative risks (RRs) from Jerrett *et al.*¹ for ozone and respiratory diseases and Burnett *et al.*⁴ for PM_{2.5} and cardiopulmonary diseases and lung cancer. Then, we apply that attributable fraction in each grid cell to future adult population (age 25 and older) and baseline mortality rates based on projections from the International Futures (IFs) integrated modelling system³². Using country-level projections per age group, we mapped and gridded to the 0.5°×0.5° grid assuming that the present-day spatial distribution of total population within each country is unchanged in the future, as well as the present-day ratio of baseline mortality for the specific causes included in the epidemiological studies and for three disease groups projected in IFs (chronic respiratory diseases, cardiovascular diseases and malignant neoplasms). We select population projections from IFs instead of those underlying RCP8.5 to ensure consistency between projections of population and baseline mortality, since the latter are not available for RCP8.5, and for consistency with Silva *et al.*¹². IFs projections of future total population are lower than those of RCP8.5 (–5% in 2030 and –27% in 2100) (Supplementary Figure 9). Had we used projections of population underlying RCP8.5, we would have likely estimated greater changes in premature mortality relative to 2000. IFs projections of baseline mortality rates reflect an aging population and regional demographic changes, showing a steep rise in chronic respiratory diseases (roughly tripling globally by 2100), particularly in East Asia and India, some regional increases in cardiovascular diseases (e.g. Middle East, Africa), and global decreases in lung cancer.

Overall uncertainty in mortality estimates includes uncertainty from the RRs and from air pollutant concentrations. First, we conduct 1000 Monte Carlo (MC) simulations separately for each model-year to propagate uncertainty from the RRs to mortality estimates. For ozone, we use the 95% Confidence Intervals (CIs) for RR reported by Jerrett *et al.*¹ and assume a normal distribution, while for PM_{2.5} we use the parameter values of Burnett *et al.*⁴ for 1000 MC simulations. Then, we calculate the average and 95% CI for the pooled results of the 1000 MC simulations for each model to quantify the spread of model results. We do not include uncertainties associated with population and baseline mortality rates, since these are not reported. As ACCMIP models used the same anthropogenic and biomass burning

emissions, we do not consider uncertainty in emissions inventories, however we acknowledge that this is an important source of uncertainty, especially in particular regions^{34–37}. Our mortality estimates are affected by our choices of and underlying assumptions regarding concentration-response functions, population, and baseline mortality rates. Although a number of factors, such as vulnerability of the exposed population and PM_{2.5} composition, vary spatially and possibly temporally, we assume that the RRs estimated for the present day apply on a global scale and in future time periods. Also, our assumption that the spatial distribution of population within each country is constant in the future likely understates the effects of rural-to-urban migration, which is currently underway and expected to continue. However, the effects of climate change on air pollutant concentrations may be somewhat spatially uniform (as opposed to changes in emissions), and the coarse grid resolution of global models would not resolve air pollutant concentrations well in urban areas.

Data Availability

Data used in this project are archived here:

Air pollutant concentrations

Atmospheric Chemistry & Climate Model Intercomparison Project (ACCMIP) datasets - <http://catalogue.ceda.ac.uk/uuid/b46c58786d3e5a3f985043166aeb862d>. Data retrieved from 08/2012 to 12/2013.

Present-day population

Oak Ridge National Laboratory (ONRL) - LandScan 2011 Global Population Dataset, <http://spruce.lib.unc.edu.libproxy.lib.unc.edu/content/gis/LandScan/>. Data retrieved on 12/05/2012.

Present-day baseline mortality

Institute for Health Metrics and Evaluation (IHME): Global Burden of Disease Study 2010 (GBD 2010) Results by Cause 1990–2010 - Country Level, Seattle, United States, 2013. <https://cloud.ihme.washington.edu/index.php/s/d559026958b38c3f4d12029b36d783da?path=%2F2010>. Data retrieved from 12/2013 to 03/2014.

Future population and baseline mortality

Web-Based IFs - The International Futures (IFs) modeling system, version 6.54., www.ifs.du.edu. Data retrieved on 07/2012.

IER model

Global Burden of Disease Study 2010. Global Burden of Disease Study 2010 (GBD 2010) - Ambient Air Pollution Risk Model 1990 – 2010. Seattle, United States: Institute for Health Metrics and Evaluation (IHME), 2013. <http://ghdx.healthdata.org/record/global-burden-disease-study-2010-gbd-2010-ambient-air-pollution-risk-model-1990-2010>. Data retrieved on 11/08/2013.

Supplementary Material

Refer to Web version on PubMed Central for supplementary material.

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Figure 1.

Impact of RCP8.5 climate change on global mortality for individual models and the multi-model average. Estimates are for 2030 and 2100 for (a) ozone respiratory mortality (9 models) and (b) PM_{2.5} IHD+STROKE+COPD+LC mortality (5 models). PM_{2.5} is calculated as a sum of species. Uncertainty for each model is the 95% CI taking into account uncertainty in RR. Uncertainty for the multi-model average is the 95% CI including uncertainty in RR and across models.

a. Ozone mortality **b. PM_{2.5} mortality**

Figure 2. Geographical impact of climate change on mortality. Estimates are for 2030 and 2100 for (a) ozone respiratory mortality and (b) PM_{2.5} IHD+STROKE+COPD+LC mortality, showing the multi-model average in each 0.5°×0.5° grid cell. PM_{2.5} is calculated as a sum of species.

**a. Ozone mortality
(deaths yr⁻¹)**

**b. PM_{2.5} mortality
(deaths yr⁻¹)**

Figure 3.

Projected mortality for ten world regions. Estimates are for 2030 and 2100 for (a) ozone respiratory mortality and (b) PM_{2.5} IHD+STROKE+COPD+LC mortality, showing the multi-model regional average. PM_{2.5} is calculated as a sum of species. Uncertainty for the multi-model regional average is the 95% CI including uncertainty in RR and across models. World regions are shown in Supplementary Figure 1.

Table 1

Sensitivity analysis for changes in global air pollution-related mortality attributable to climate change. Estimates are for multi-model averages (deaths yr⁻¹) for the deterministic results.

	PM _{2.5} -related mortality		Ozone-related mortality	
	2030	2100	2030	2100
Base results	56,300	218,000	10,700	128,000
PM _{2.5} using Krewski et al. ²	66,200	318,000	–	–
Present-day (2011) population	35,500	93,800	2,970	59,400
Present-day (2010) baseline mortality rates	69,600	510,000	2,790	13,300
Present-day population and baseline mortality rates	43,300	144,000	2,300	14,500