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## Ensemble projections of wildfire activity and carbonaceous aerosol concentrations over the western United States in the mid-21st century

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### Abstract

We estimate future wildfire activity over the western United States during the mid-21<sup>st</sup> century (2046–2065), based on results from 15 climate models following the A1B scenario. We develop fire prediction models by regressing meteorological variables from the current and previous years together with fire indexes onto observed regional area burned. The regressions explain 0.25–0.60 of the variance in observed annual area burned during 1980–2004, depending on the ecoregion. We also parameterize daily area burned with temperature, precipitation, and relative humidity. This approach explains ~0.5 of the variance in observed area burned over forest ecoregions but shows no predictive capability in the semi-arid regions of Nevada and California. By applying the meteorological fields from 15 climate models to our fire prediction models, we quantify the robustness of our wildfire projections at mid-century. We calculate increases of 24–124% in area burned using regressions and 63–169% with the parameterization. Our projections are most robust in the southwestern desert, where all GCMs predict significant ( $p < 0.05$ ) meteorological changes. For forested ecoregions, more GCMs predict significant increases in future area burned with the parameterization than with the regressions, because the latter approach is sensitive to hydrological variables that show large inter-model variability in the climate projections. The parameterization predicts that the fire season lengthens by 23 days in the warmer and drier climate at mid-century. Using a chemical transport model, we find that wildfire emissions will increase summertime surface organic carbon aerosol over the western United States by 46–70% and black carbon by 20–27% at midcentury, relative to the present day. The pollution is most enhanced during extreme episodes: above the 84<sup>th</sup> percentile of concentrations, OC increases by ~90% and BC by ~50%, while visibility decreases from 130 km to 100 km in 32 Federal Class 1 areas in Rocky Mountains Forest.

### Keywords

wildfire; ensemble projection; fuel load; aerosol concentration

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## 1. Introduction

Wildfire emissions can adversely affect air quality locally and downwind (e.g., Wotawa and Trainer, 2000; Val Martin et al., 2006). Wildfire activity in North America is strongly related to weather conditions, such as temperature and humidity (e.g., Westerling et al., 2002; Gillett et al., 2004). Observations show an increasing trend in the area burned by fires in the western United States, suggesting that climate change may have already enhanced wildfire activity in this region (Westerling et al., 2006). In this study, we use two different fire prediction schemes together with an ensemble of climate models to estimate area burned in the western United States at midcentury under a moderately warming scenario. With a chemical transport model, we then calculate the consequences of changing wildfire activity for air quality across the West.

Meteorological conditions greatly influence the extent of area burned by fires, whether they are started by lightning or by human activity (e.g., Flannigan et al., 2009; Littell et al., 2009). Higher temperatures, lower precipitation, and/or lower relative humidity promote larger fires (Flannigan and Harrington, 1988). In addition to current weather, meteorological conditions during the preceding months or years may also affect area burned by influencing the amount of fuel as well as fuel moisture (e.g., Westerling et al., 2002; Crimmins and Comrie, 2004). In some regions of the western U.S., land use management may help reduce wildfire severity (Prichard et al., 2010).

Previous studies investigating the impact of future climate change on wildfire activity in North America have projected increases in area burned during the 21<sup>st</sup> century, but yield a wide range of results, as shown in Table S1. Included in the Table are studies following various scenarios, since the area burned predictions are sensitive to the scenario as well as to the prediction tool and the choice of climate model. These studies generally relied on projections from only 1–3 climate models, so could not explore the model dependence of their results. For example, to project area burned in Canada, Flannigan et al. (2005) used stepwise regression to build relationships between observed area burned and a suite of meteorological variables and fire indexes. Applying these relationships to two climate models, they found a doubling of area burned by 2100. Using a similar approach but only one climate model, Balshi et al. (2009) found that area burned across Alaska and Canada would double by 2050. Again, with a similar approach and one climate model, Spracklen et al. (2009) predicted an increase of ~50% in area burned over the western United States by the midcentury, relative to present day. The discrepancies among these studies are likely caused by differences in fire prediction schemes, the model sensitivities to changing greenhouse gases, and the choice of climate scenarios.

Most studies in Table S1 used regression methods, such as linear or exponential regressions or the Multivariate Adaptive Regression Spline (MARS) method, to predict area burned, with the regressions developed by fitting time series of observed area burned (e.g. Flannigan et al., 2005; Balshi et al., 2009). In recent years, new parameterizations or process-based fire models implemented in Dynamic Global Vegetation Models (DGVM) have been developed to simulate present-day fire activity on both global and regional scales (Crevoisier et al., 2007; Pechony and Shindell, 2009; Thonicke et al., 2010). These studies rely on empirical functions for fire activity based on variables such as soil moisture, temperature, relative humidity, precipitation, or road density. Each method has advantages and limitations, as we discuss in a short review in section A of the supporting information (SI).

In this study, we investigate future wildfire activity over the western United States during the mid-21<sup>st</sup> century (2046–2065), using two different fire prediction approaches and taking advantage of output from an ensemble of climate models from the World Climate Research

Programme's (WCRP's) Coupled Model Intercomparison Project phase 3 (CMIP3) multi-model dataset (Meehl et al., 2007). Our use of a model ensemble allows us to project future fire activity with greater confidence than previous studies. We focus on the metric of area burned because calculation of fire emissions requires this variable. As in Spracklen et al. (2009, hereinafter S2009), we develop fire models by regressing meteorological variables and fire indexes onto observed area burned. Here we improve on S2009 by allowing for the dependence of fires on meteorology in prior years as well as in the current year. We also develop and evaluate a fire parameterization to test the robustness of predictions from the regression approach. We examine the fire-induced changes in organic carbon (OC) and black carbon (BC) aerosol concentrations at mid-century using the GEOS-Chem chemical transport model (CTM) driven by the Goddard Institute for Space Studies General Circulation Model 3 (GISS GCM3). An illustration of these steps is shown in Fig. 1.

## 2. Fire model development

### 2.1. Development and evaluation of regressions

We divide the western United States ( $31^{\circ}$ – $49^{\circ}$ N,  $101^{\circ}$ – $125^{\circ}$ W) into six ecoregions (Fig. 2), aggregated from the 18 ecosystems defined by Bailey et al. (1994). The Pacific Northwest and Rocky Mountains Forest regions are mainly mixed and/or coniferous forests. The more arid California Coastal Shrub, Desert Southwest, and Nevada Mountains/Semi-desert are dominated by shrubland, savanna, and desert. The Eastern Rocky Mountains/Great Plains region is mainly grassland. We develop relationships between the observed area burned and surface meteorological variables and fire indexes using a stepwise regression method in each ecoregion. The approach follows that described in S2009 except for three differences. First, we adopt new meteorological datasets which are more complete in space and time compared with the one used in S2009. Second, we include predictors from the current and previous years in the regression. Third, we improve the statistical method in selecting terms in the regression.

As in S2009, we use a gridded dataset ( $1^{\circ} \times 1^{\circ}$ ) of monthly area burned derived from fire reports for 1980–2000 (Westerling et al., 2002) and extended to the year 2004. The gridded area burned for the entire fire season (May–October) is binned into ecoregions as the predictands in the regression. Meteorological observations and fire indexes are used to calculate the predictors. We use monthly-averaged mean and maximum temperature ( $T$  and  $T_{max}$ ) and total monthly precipitation ( $Prec$ ) from the United States Historical Climatology Network (USHCN), a dataset that includes meteorological fields from 384 stations in the western United States (Easterling et al., 1996). We also use seven fire indexes from the Canadian Fire Weather Index system (CFWIS, Van Wagner, 1987), which are calculated using daily variables from 234 sites in western U.S. provided by the Global Surface Summary of the Day (GSOD, see section B of the SI). Since the USHCN sites do not report  $RH$ , we compute monthly mean  $RH$  from the daily values derived from the GSOD data. The four monthly meteorological fields ( $T$ ,  $T_{max}$ ,  $Prec$ ,  $RH$ ) are used to calculate seasonal (winter, spring, and summer), annual, and fire-season means. The seven daily fire indexes are used to calculate mean and maximum values during the fire season. This yields 34 factors for the current year. We account in part for the impact of weather on fuel loading by including meteorological factors from the previous 1–2 years in the regressions, leading to 102 potential predictors for the stepwise regression for each ecoregion.

We develop an empirical relationship between observed area burned (1980–2004) and meteorological variables and fire indexes, using linear forward stepwise regression in each ecoregion. To select a factor as a predictor, we require two criteria. First, the chosen factor must have the largest contribution to the  $F$  value of the predictand among unselected factors. Second, the factors selected must be independent of each other, which we test for using a  $p$

value  $> 0.5$  for their correlation, a weak test for independence. These two criteria rule out many possible predictors that were selected in previous studies (e.g., Flannigan et al., 2005; Littell et al., 2009, and S2009), but they increase the stability of the regressions (Philippi, 1993).

Table 1 shows the regression relationships for each ecoregion, and Fig. 3 compares the model fits to observations. The regressions explain 25–60% of the variance in area burned, similar to the results of S2009. The regressions best fit the observed area burned time series ( $R^2 \approx 0.6$ ) in forested ecosystems (Pacific Northwest and Rocky Mountains Forest). Additional statistical test shows that such fits are robust (see section C of the SI). In these ecosystems, current meteorology is most important, as hot weather and low rainfall dry out the plentiful fuels. The fit for the Desert Southwest is slightly less than that in S2009,  $R^2 = 0.46$  compared to 0.49, but both approaches select  $T_{max}$  as the main predictor. By allowing for terms from the previous year, the fit for the Nevada Mountains/Semi-desert increases from  $R^2=0.37$  in S2009 to  $R^2=0.52$  in this study. In this region,  $RH$  in the previous fire season is the most important term since the spread of large fires depends on moisture during the previous year enhancing the fuel load (Crimmins and Comrie, 2004). The regression model performs more poorly for the California Coastal Shrub ecoregion than in S2009, even allowing for lagged relationships.

## 2.2. Development and evaluation of fire parameterization

We develop a physical parameterization for area burned over the western United States, building on the approach of Crevoisier et al. (2007) for boreal fire probability, and the form of the empirical function in Pechony and Shindell (2009) for global flammability. We use the 32 km gridded data from the North American Regional Reanalysis (NARR) (Mesinger et al., 2006), which we aggregate onto the same  $1^\circ \times 1^\circ$  grid as area burned data. We use monthly fire and climate data from 1980 to 2004, and select those grid points with area burned larger than 10 ha, yielding 13228 cells. These include 13.3% of the total land grid cells but 99.8% of the total area burned. We bin these fire cells by three meteorological variables (surface air temperature, relative humidity, and precipitation) and calculate the probability of fire cells and the averaged logarithmic monthly area burned ( $\ln(AB)$ , ha) within each bin.

As Fig. 4a shows, fire cells generally have higher temperature and lower relative humidity and precipitation. More than 75% of fire cells have temperatures higher than  $15^\circ\text{C}$ . Higher temperatures generally increase the probability of fire, but for temperatures exceeding  $23^\circ\text{C}$ , fire becomes less likely (red points, 27% of fire cells), due either to lack of fuel (e.g. desert) or to precipitation from the North American monsoon. More than 85% of the fire cells show an inverse relationship with relative humidity (blue points in Fig. 4b). For some extremely dry regions ( $RH < 25\%$ ) fire spread is limited by fuel availability and fires are uncommon (red points). There is an inverse relationship between precipitation and fires, and almost all fires occur when the monthly rainfall is  $< 2.5 \text{ mm day}^{-1}$  (Fig. 4c). As in Crevoisier et al. (2007), we found no useful information for surface wind speed.

Based on the distribution plots, the relationships used in the parameterization of Pechony and Shindell (2009), and trial-and-error, we propose a parameterization for the monthly area burned ( $AB$ , ha) as follows,

$$\ln(AB) \propto P(T, RH, R) = F(T) \times G(RH) \times H(R) \quad (1)$$

where  $T$  is the monthly mean temperature ( $^\circ\text{C}$ ),  $RH$  is the monthly mean relative humidity (%), and  $R$  is the monthly average precipitation ( $\text{mm day}^{-1}$ ) from the NARR. Here  $F(T) = T$

$T_b$ ,  $G(RH) = (1 - RH/100)^2$ , and  $H(R) = 1/(R+c)$  with  $T_t = 15^\circ\text{C}$  and  $c = 1 \text{ mm day}^{-1}$ . The relationship between  $\ln(AB)$  and the different terms of Equation (1) are shown in Figs. 4e–g. These functions predict approximately linear relationships of area burned for each term. The relationship between  $\ln(AB)$  and  $P(T, RH, R)$  is linear if  $P < 1$  (Fig. 4h), which has a probability >95% (Fig. 4d).

We extend Equation (1), which is based on monthly data, to a daily scale with two assumptions. First, we choose  $c = 0.2 \text{ mm day}^{-1}$  for  $H(R)$  in the daily calculation. Evaluation shows that this choice results in <1% change in the long-term mean values relative to the results with  $c = 1.0 \text{ mm day}^{-1}$ . However, it provides a sharp contrast in area burned between days with and without precipitation. Second, we assume that the distribution of orography and the availability of fuels are similar within an ecosystem, and use the 18 ecosystems defined by Bailey et al. (1994) in the western United States. To reflect the impact of factors other than meteorology, such as elevation and fuel load, we introduce a fire potential coefficient  $a$  into the parameterization. With these two assumptions, we propose a data-constrained parameterization for daily area burned over the western United States as follows:

$$\ln(AB) = \begin{cases} \frac{\alpha \cdot T \cdot (1 - RH/100)^2}{T_t \cdot (R + 0.2)} & \text{if } T > T_t \text{ and } R < R_t \\ \text{no fire} & \text{otherwise} \end{cases} \quad (2)$$

We determine  $a$  for each ecosystem such that the long-term annual mean area burned matches that observed. Two threshold values  $T_t = 15.0^\circ\text{C}$  and  $R_t = 2.5 \text{ mm day}^{-1}$  (see Figs. 4a and 4c) are employed to remove noise from the predicted area burned. We use daily NARR meteorology as input for equation (2). We do not apply the NARR dataset for regressions, because it shows some biases for precipitation, relative humidity, and surface wind speed, which would affect the calculated FWI indexes. For example, wind speed in the NARR does not correlate with the GSOD site observations, while the NARR relative humidity at values above 0.5 correlates only weakly ( $R < 0.6$ ).

We evaluate the interannual variability in area burned calculated by the parameterization in Fig. 5. Compared with the regression method (Fig. 3), the parameterization performs less well for the Pacific Northwest ( $R^2 = 0.47$ ), Desert Southwest ( $R^2 = 0.28$ ), and Rocky Mountains Forest ( $R^2 = 0.49$ ), but performs better for the Eastern Rocky Mountains/Great Plains ( $R^2 = 0.69$ ). In these three ecoregions, the current year's meteorology dominates (Table 1). The parameterization fails in the California Coastal Shrub and Nevada Mountains/Semi-desert ecoregions, where the regression method finds that the previous year's weather strongly influences the area burned. After excluding the predictions from these two ecoregions, the correlation coefficient between simulation and observations is 0.71 for the monthly data and 0.57 for the deseasonalized time series over the western U.S. during 1980–2004. Additional evaluations show that the parameterization reproduces the spatial pattern and the seasonality of area burned in most ecoregions (see section D of the SI).

### 3. Ensemble projection of future area burned

We use both the regressions and parameterization with future climate from 15 GCMs to project future area burned in the mid 21<sup>st</sup> century. We assume that the relationships we derived between area burned and meteorological factors are the same in the future as at present. We use daily output from the NASA/GISS Model 3 (Hansen et al., 2002) and from 14 other GCMs in the CMIP3 archive (Meehl et al., 2007) (Table 2). The simulated meteorology includes daily mean and maximum temperature, total precipitation, and surface wind speed. We calculate daily  $RH$  for the CMIP3 models from other archived



meteorological variables. Output from the Intergovernmental Panel on Climate Change (IPCC) 20C3M scenario is used for the present day (1981–2000). For this scenario, the models are freely running but forced by the well-mixed greenhouse gases increasing as observed in the 20<sup>th</sup> century. The resulting model meteorology matches the observed, but only in a decadal average sense. Output from the A1B scenario is used for the future (2046–2065). This scenario assumes significant innovations in energy technologies, which place more emphasis on non-fossil fuels, improve energy efficiency, and reduce the cost of energy supply, resulting in decreased CO<sub>2</sub> emission after 2050. The 522 ppm of CO<sub>2</sub> projected by 2050 in the A1B scenario is consistent with two moderate scenarios in the new Representative Concentration Pathways (RCPs) developed by the IPCC, RCP 6.0 and RCP 4.5 (Moss et al., 2010).

When applying the regressions to predict fires, we merge the monthly gridded output from each GCM into the six ecoregions and scale them using the mean observations for 1981–2000 from the USHCN. However, we scale the daily variables with the GSOD to calculate CFWIS indexes and monthly *RH*. The regressions predict some negative burned areas and we set these to zero. For the prediction with the parameterization, we interpolate the CMIP3 data with different spatial resolutions (Table 2) onto the same 1°×1° grid using the inverse distance weighted method, as in other studies (e.g. Balshi et al., 2009). The regridded daily GCM outputs are then scaled to the monthly mean NARR reanalyses of 1981–2000 for each grid box. Our approach does not resolve the impacts of topography on meteorology as well as the use of statistically downscaled GCM data might (Maurer et al., 2007), but CMIP3 downscaled datasets (e.g., <http://gdo-dcp.ucllnl.org/>) lack variables (*RH* and wind speed) needed for our prediction methods. The results given in this study for each ecoregion are shown as the median value of the ensemble of predictions from 15 GCMs and median value of the 15 ratios of predicted to observed area burned. We choose to use the median rather than the mean values to avoid skewing by extreme model results.

Prediction of present-day area burned with GCM meteorology reproduces observed values in most ecoregions for both regressions and the parameterization (Tables 3 and 4, and section E in the SI). In response to the warmer and drier summer climate (see section F in the SI), the median predicted area burned from the regressions increases in every ecoregion at the midcentury relative to present day, as shown in Table 3, and by Fig. 6a and Fig. 7a. The last column of Table 3 shows the number of models that predict a ratio within the range of ±20% of the median ratio, so higher numbers indicate greater agreement among the projections. The predicted time series for present-day area burned in Fig. 6a are much smoother than observed, as they show the median result of the 15 GCMs each year. The interannual variation is much greater for the individual models, and some of them predict a 30% higher standard deviation than do the observations. The regression method projects increases in annual area burned of 24–124% depending on ecoregion, with the largest median ratio of future to present-day area burned in the Desert Southwest (Fig. 7a). Table 3 and Fig. 7c show that only a few GCMs predict a significant increase in area burned for the Pacific Northwest and California Coastal Shrub, while most predict an increase in the desert and semi-desert regions. Ratios of future to present-day area burned for individual GCMs are shown in Fig. 7, and we discuss this figure below.

Table 5 shows the changes in regression terms and their contributions to the projected increases in area burned from the regressions. The Table makes clear the importance of considering variables other than just temperature when considering future fire activity. Changes in fuel moisture codes and temperature contribute the most to the change in area burned, and more GCMs predict significant changes in these factors than in relative humidity and rainfall. For the ecoregions where the changes in area burned are dominated by fuel moisture codes or temperature, we find that the GCMs that predict significant increases

in those factors also predict significant increases in area burned (Tables 3 and 5). In two regions, the Pacific Northwest and California Coastal Shrub, the increases in area burned related to drought codes are offset to some extent by changes in relative humidity. Only a few GCMs predict significant increases in area burned in these regions (Table 3 and Fig. 7c).

The projections of future area burned with the parameterization are shown in Table 4 and Fig. 6b for the four ecosystems whose area burned is captured for the present-day using this approach. The parameterization yields increases of 63–169% in annual area burned, with the largest increases in the Pacific Northwest and the Rocky Mountain Forest (Table 4 and Fig. 7b). The parameterization method generally gives a higher ratio of future area burned to present day than does the regression approach, and the spread of area burned predicted by the 15 GCMs is generally larger (Figs. 6, 7) because of the exponential relationship.

Because temperature plays a major role in the parameterization, more GCMs predict significant increases in future area burned for forested ecoregions with this approach than with the regressions (Fig. 7d). As shown in Fig. S5, the projected median increases in summer temperature are greater than 2 °C in all ecoregions, and the decreases in relative humidity and precipitation are around 1% and 0.1 mm day<sup>-1</sup>, respectively. We explored the regional sensitivity of area burned to changes of this magnitude by first applying incremental changes to the NARR observed time series of temperature, relative humidity, and precipitation in each ecoregion and then recalculating area burned (Fig. 8). As expected, warmer temperatures and lower humidity and rainfall cause increases in area burned, and the responses to the first two are symmetric with respect to the sign of the change. The effect of the increase in rainfall is artificial, as it leads to daily rainfall and almost no burning. The results in Fig. 8 suggest that the changes in temperature are the primary driver of the increase in area burned, and that the two forested regions (Pacific Northwest and Rocky Mountains Forest) are most sensitive to increases in temperature.

Both the regression and the parameterization methods sometimes yield unrealistic projections of fire activity (Fig. 7). As noted above, predictions of negative area burned by the regressions are set to zero. Using the parameterization method, area burned increases by unreasonably large factors (10–70) with output from two GCMs in the Pacific Northwest, Rocky Mountain Forest, and Eastern Rocky Mountains/Great Plains. These same two models project large increases in temperature and large decreases in precipitation in summer for these regions, relative to other models (Fig. S5). Four other models give increases of factors of 6–15 for the Rocky Mountain Forest, again due to the relatively large changes in temperature and precipitation there. Except for these extreme cases, most models give ratios of 1–4 for future/present area burned in all ecoregions with either the regression or parameterization method. Predictions with the regressions show the largest uncertainty in the Eastern Rocky Mountains/Great Plains (Fig. 7a), because of the large spread of changes in meteorological variables that all contribute to increases in area burned (Fig. S5; Table 1). However, projections with the parameterization exhibit the largest uncertainty in the Rocky Mountains Forest (Fig. 7b), due to the large spread in the meteorological changes here as well and the high sensitivity of area burned to these changes.

An advantage of the parameterization method is that it predicts the seasonality of fires. Relative to present day, fires show larger area burned during the peak months. For example, area burned increases by 65% in August in the Pacific Northwest, 68% in June in Desert Southwest, 390% in August in Rocky Mountains Forest, and 190% in August in Eastern Rocky Mountains/Great Plains by the midcentury (Fig. 9). The probability of large fires, defined as the average percentage of grid cells that have daily area burned >100 ha during

the fire season, is small for the present-day, 0.3% to ~2% for the four ecosystems, and increases by factors of 2 to 3 with the parameterization.

We examine the predicted length of fire season, which we define as the period when the daily area burned is larger than 100 ha in at least one grid box ( $1^{\circ} \times 1^{\circ}$ ). In response to the changing climate, the median start date changes from May 12 to April 28 and the median end date from October 4 to October 14, so that the median length of the fire season increases by over 3 weeks or 16% (Fig. 10). This change is robust, with 14 out of 15 GCMs projecting significant increases in length ( $p < 0.05$ ). In most models, the fire season begins in the Desert Southwest and ends in the Pacific Northwest in both the present-day and future atmospheres. The median length of the period with large fires ( $>100$  ha) increases by 20–42% in the four ecoregions, with the largest extension in the Eastern Rocky Mountains/Great Plains. Our results are consistent with the extension of fire season in boreal and temperate areas as projected by previous studies (Wotton and Flannigan, 1993).

#### 4. Estimate of future biomass burned

Fuel consumption is the amount of both live and dead biomass burned per unit area. We calculate present-day fuel consumption based on the fuel load database of the United States Forest Service, the Fuel Characteristic Classification System (<http://www.fs-fed-us/pnw/fera/fccs/-downloaded-on-Ma-2011> FCCS, <http://www.fs.fed.us/pnw/fera/fccs/>, downloaded on May 12<sup>th</sup>, 2011) (McKenzie et al., 2007; Ottmar et al., 2007). The 1 km  $\times$  1 km fuelbed map for the conterminous U.S. is derived from the distribution of vegetation types from the Landscape Fire and Resource Management Planning Tools (LANDFIRE, <http://www.landfire.gov/>). This is a more complete version of the FCCS than available at the time of the S2009 study. We calculate fuel consumption following Table 4 of S2009. As in S2009, we assume that fires burn with 25% high, 25% medium, and 25% low severity, with 25% of the area unburned. An updated map of fuel consumption is shown in Fig. 11a. The value ranges from 0–1 kg dry mass (DM)  $m^{-2}$  in arid and semi-arid southwestern area. However, in northern forest region, the fuel consumption is as high as 9.5 kg DM  $m^{-2}$ . We acknowledge that using the fixed fuel consumption map does not reflect the impact of interannual variability of biomass on fire emissions. We assume that fuel consumption throughout the western U.S. remains constant for 50-year period, because the simulation with a DGVM shows small changes in the cover of vegetation with a high fuel load (see section G in the SI).

The annual mean biomass burned (Fig. 11b and Table 6) is the product of gridded area burned and fuel consumption. The largest values of present-day biomass burned are located in forest ecoregions, with 30% in the Pacific Northwest and 39% in the Rocky Mountains Forest. Although the Nevada Mountains/Semi-desert ecoregion has the largest area burned (35% of the total), it contributes only 9% to the total biomass burned because of low fuel availability (Table 6). The annual total biomass burned by wildfire is 13.0 Tg during 1980–2004 over the western United States. This is smaller than the estimate in S2009, 14.2 Tg, because of the updated FCCS fuel bed map.

To obtain the GCM estimates of biomass burned, we combine the projected area burned that has been allocated to the  $1^{\circ} \times 1^{\circ}$  grid (see section H of the SI) with the fuel consumption. As shown in Fig. 11c, the present-day biomass burned calculated with the median area burned predicted by the regressions resembles that based on the observed area burned (Fig. 11b). The prediction overestimates biomass burned in the Rocky Mountains Forest (especially in Montana) by as much as 70%. The predicted total biomass burned over the western United States for the regression method is 17.0 Tg  $yr^{-1}$  for 1986–2000 (Table 6), higher than the value based on observed area burned, 13.0 Tg  $yr^{-1}$ . The parameterization predicts slightly



smaller present-day biomass burned (Fig. 11d), 12.1 Tg yr<sup>-1</sup>. These differences reflect the discrepancies between modeled and observed area burned for the two methods (Tables 3 and 4).

The median values of future area burned predicted by the two approaches are used to calculate future biomass burned as shown in Figs. 11e and 11f and Table 6. Both methods predict similar spatial patterns and larger amounts of biomass burned at midcentury, with the greatest increases in the Pacific Northwest and the Rocky Mountain Forest. The annual total biomass burned over the western United States increases by 77% with the regressions and 140% with parameterization.

## 5. Simulations of carbonaceous aerosols

We calculate the future fire emission based on the predicted biomass burned and the emission factors from Andreae and Merlet (2001). With the median area burned predicted by the regressions, we estimate that the total wildfire emission is 133 Gg C yr<sup>-1</sup> for OC and 9 Gg C yr<sup>-1</sup> for BC over the western United States for the present day, somewhat larger than the values calculated with observed area burned, 95 Gg C yr<sup>-1</sup> for OC and 7 Gg C yr<sup>-1</sup> for BC, due to the overestimate of present-day area burned in Rocky Mountains Forest by regressions (Table 3). The predicted emissions increase by ~80 % for both OC (to 238 Gg C yr<sup>-1</sup>) and BC (to 16 Gg C yr<sup>-1</sup>) by midcentury. Results from the parameterization give present-day emissions of 88 Gg C yr<sup>-1</sup> for OC and 6 Gg C yr<sup>-1</sup> for BC. These emissions increase by ~170 % for OC (to 235 Gg C yr<sup>-1</sup>) and ~150 % for BC (to 16 Gg C yr<sup>-1</sup>) by the midcentury.

To simulate the distributions of BC and OC over the western United States in the present-day and future atmospheres, we use the GEOS-Chem CTM (v08.03.01). The CTM, driven with present-day and future meteorological fields from the GISS GCM3, has been used to estimate the impact of changing climate on the future air quality in the United States (e.g., Wu et al., 2008; S2009). The model set-up is similar to that described in S2009, and we note any differences below.

The CTM transports hydrophobic and hydrophilic BC and OC. Hydrophobic aerosols become hydrophilic with an e-folding time of 1.2 days (Cooke et al., 1999). Global anthropogenic and biofuel emissions for BC and OC are from Bond et al. (2007); the emissions over North America are imposed with the seasonality following Park et al. (2003). Outside the western United States, climatological biomass burning emissions are from Lobert et al. (1999), with seasonality from Duncan et al. (2003). The main biogenic sources for OC are terpenes, which are calculated with the Model of Emissions of Gases and Aerosols from Nature (MEGAN), version 2.1 (Guenther et al., 2006). We assume a 10% yield of hydrophilic OC from these emissions. To isolate the effects of changes in fires in this study, we use constant emissions from anthropogenic and biofuel emissions and from biomass burning outside the western United States. Biogenic emissions vary in response to the changing climate.

We conduct three simulations for the present day (1997–2001) and for midcentury (2047–2051), with a six month spin-up, as shown in Table 7. These include simulations with monthly fire emissions from the median of GCMs predictions with the regression (Reg) and the parameterization (Param), and one with daily emissions from the latter (Param.D). We first aggregate the 1°×1° calculated fire emissions onto the 4°×5° grids of the CTM. Since our use of median values smoothes the interannual variability of area burned, we use the same total amount of area burned in a specific month in a given simulation each year, but the locations of fires are redistributed every year with the random method described in H of the SI.

We first examine our present-day simulations of carbonaceous aerosol using the median wildfire emissions from the 15 GCMs. With the regression method, the CTM simulates highest concentrations for OC,  $\sim 2.0 \mu\text{g m}^{-3}$ , in Oregon, Idaho, and western Montana during summer in the present (Fig. 12a). The BC concentrations are relatively high,  $\sim 0.2 \mu\text{g m}^{-3}$ , in these regions as well (Fig. 12c), but high values are also predicted over the coastal states without large fire emissions, because the dominant fraction ( $>70\%$ ) of BC emission is from fossil fuel combustion (Park et al., 2003). The CTM driven by fire emissions from the parameterization gives similar present-day distributions of OC (Fig. 12b) and BC (Fig. 12d) compared with those from the regressions, but with smaller magnitudes due to the smaller fire emissions.

At mid-century using both approaches, the simulated OC and BC fields show large increases in summer, relative to the present-day. With the regression method, the mean summertime concentration of OC aerosol increases from  $0.96 \mu\text{g m}^{-3}$  in the present-day to  $1.40 \mu\text{g m}^{-3}$  in the future (46%), while mean BC increases from  $0.14$  to  $0.17 \mu\text{g m}^{-3}$  (20%). The increase is most prominent over Oregon, Idaho, western Montana, and Wyoming ( $1.15 \mu\text{g m}^{-3}$  for OC and  $0.07 \mu\text{g m}^{-3}$  for BC, Figs. 12i and k). Similar results are obtained in the simulation using the parameterization method (Figs. 12j and l), which predicts a 70% increase of mean OC aerosol from  $0.89 \mu\text{g m}^{-3}$  in the present-day to  $1.41 \mu\text{g m}^{-3}$ . BC for this method increases 27%, from  $0.13 \mu\text{g m}^{-3}$  to  $0.17 \mu\text{g m}^{-3}$ . We calculate the greatest change in Idaho, where OC increases by  $1.59 \mu\text{g m}^{-3}$  and BC by  $0.09 \mu\text{g m}^{-3}$ . In the simulations for mid-century, biogenic OC emissions increase by  $30 \text{ Gg C yr}^{-1}$  (19%), due to higher temperatures. However, this increase is much smaller than the increase in fire emissions from the regression and parameterization approaches,  $105 \text{ Gg C yr}^{-1}$  and  $148 \text{ Gg C yr}^{-1}$  respectively. More than 77% of the changes in OC may be attributed to the changes in wildfire emissions, similar to the estimate by S2009.

Using the simulations with the parameterization of area burned, we examine the impact of climate change on the frequency of extreme pollution events from wildfires in 32 Federal Class 1 areas (<http://www.epa.gov/>) in the Rocky Mountains Forest ecoregion as shown in Fig. 13. Values at the high end of the distributions represent pollution episodes. We find that future wildfire activity would increase aerosol levels dramatically on pollution days. For days characterized by aerosol levels above the 84<sup>th</sup> percentile (i.e., the most polluted 30 days during the fire season), OC increases on average by  $\sim 90\%$  ( $1.55 \mu\text{g m}^{-3}$ ) and BC increases by  $\sim 50\%$  ( $0.09 \mu\text{g m}^{-3}$ ). Following the method in Park et al. (2006) we estimate that such enhancement in air pollutants increases the average haze index from 10.9 deciviews to 13.5 deciviews ( $> 84^{\text{th}}$  percentile, Fig. 13c), which corresponds to a change in visibility from 130 km to 100 km based on the definition in Pitchford and Malm (1994). Our predictions are conservative since we use daily emissions based on the median area burned.

## 6. Discussion and Conclusions

We estimated the changes in future wildfire activity and their impact on carbonaceous aerosols over the western United States during the mid-21<sup>st</sup> century (2046–2065). We developed both regressions and a parameterization as fire prediction tools and used an ensemble of climate model projections for 2050 conditions together with the CTM GEOS-Chem. Use of this ensemble allowed us to predict trends in future fire activity with greater confidence than previous studies, and to diagnose the consequences of these trends for air quality across the western United States. We list the key results for our study below, followed by a more detailed summary of our results.

- The median model results show consistent increases of 24–169% in area burned across much of the western United States in the warmer and drier future atmosphere, regardless of projection method.

- For the regressions, most GCMs show significant increases in area burned in the Desert Southwest, Nevada Mountains/Semi-desert, Rocky Mountains Forest, and Eastern Rocky Mountains/Great Plains. The projections are less robust in the Pacific Northwest and California Coastal Shrub ecoregions.
- For the parameterization, most GCMs predict significant increases in area burned in the Pacific Northwest, Desert Southwest, Rocky Mountains Forest, and Eastern Rocky Mountains/Great Plains.

Both regressions and the parameterization best capture the observed interannual variability of wildfires over forested regions. In addition, the parameterization successfully represents the seasonality of area burned in most regions of western United States. In response to the warmer and drier summer climate in the A1B scenario, the regressions project that the annual area burned will increase by 24–124%, with the largest increases in the Desert Southwest where the area burned is especially sensitive to temperature. Under the same climate scenario, the parameterization projects increases of 63–169% in annual area burned, but with the largest relative increases in the Pacific Northwest and the Rocky Mountain Forest. Using the parameterization, we showed that the length of the fire season increases by more than three weeks in the future atmosphere. These changes in fire emissions worsen air quality by the midcentury. Using the area burned from regressions, GEOS-Chem simulations predict that the summertime mean surface aerosol concentrations will increase by 46% for OC and 20% for BC during 2047–2051 relative to the values in 1997–2001. With the parameterization, these changes are 70% and 27% respectively. The intensified wildfire activity has a larger impact on the extreme events.

To our knowledge, this is the first time an ensemble of climate models has been applied to present-day relationships between meteorology and area burned to calculate the effect of changing climate on the area burned by wildfires and the resultant pollution. Our use of an ensemble provides confidence in our main results, that fire activity will increase in the future atmosphere, with potentially serious consequences for air quality. Although the climate models show large variation in their projections of key variables associated with fire weather, application of a multi-model approach allows us to overcome the challenges posed by these variations by focusing on median results and the number of GCMs that give significant changes in area burned.

Our use of a multi-model ensemble allowed us not only to identify robust results, but also to quantify the uncertainty in our projections and to diagnose model outliers. While all models project significant increases in surface temperature averaged over the West, the spread in model response is several degrees. Trends in precipitation and relative humidity, on the other hand, are not robust across models for the region. The western U.S. lies in a transition zone between the low latitudes, where most models project drier conditions in a future atmosphere, and higher latitudes, where most models predict a more moist climate (Christensen et al., 2007). These differences among climate models in the projections of future temperature, precipitation, and relative humidity lead to substantial variation among the projections for area burned, particularly for the parameterization approach. By examining the median response in fire activity to changing climate conditions, we discounted the results from outlier models in our study.

Wildfire is a consequence of the interactions between climate, people, and fuels. In this study we focus only on the impact of climate change alone because this factor has the dominant impact on the size, timing and severity of fires (Bessie and Johnson, 1995). However, other factors could also influence future wildfires. For example, decades of effective fire suppression in the West has promoted the unnatural accumulation of forest

fuels and may increase the probability of unprecedentedly large fires in the future (Schoennagel et al., 2004), an effect we do not take into account.

Given this limitation, our study shows that wildfire activity will likely increase in the western U.S. in the warmer A1B climate at mid-century. Our regression approach shows area burned doubling in the Desert Southwest by 2050, while our parameterization predicts area burned doubling across the Pacific Northwest, Rocky Mountains Forest, and the Eastern Rockies/Great Plains. These increases in area burned together with the longer fire season calculated in our study could significantly limit visibility in parks and wilderness areas and worsen regional air quality.

## Supplementary Material

Refer to Web version on PubMed Central for supplementary material.

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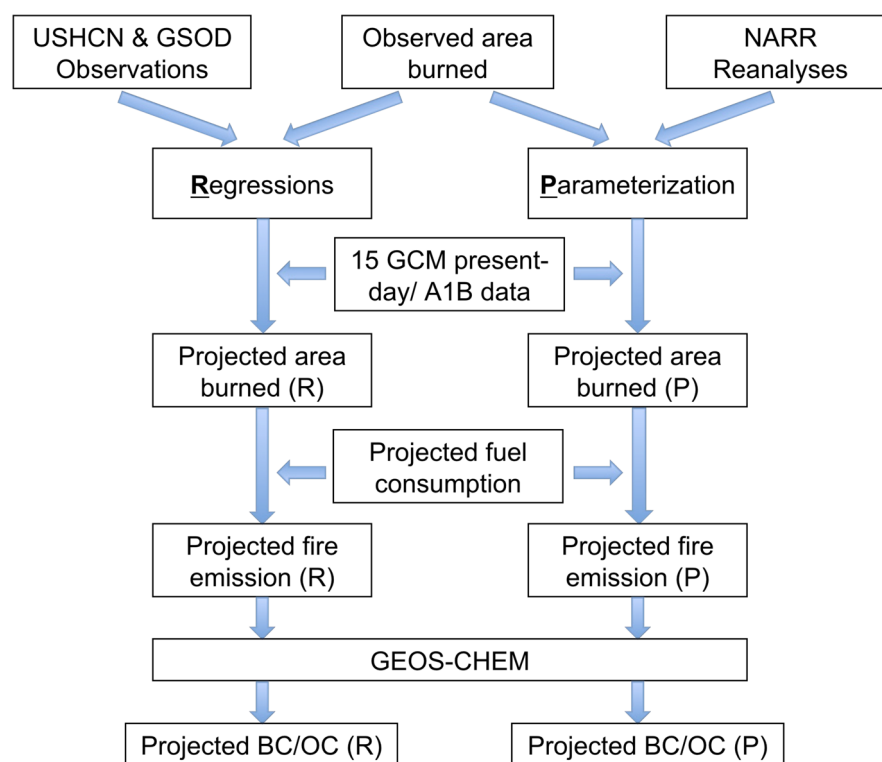
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### Highlights

1. We apply ensemble projection for future wildfire using output from 15 GCMs
2. We develop and evaluate both regressions and a parameterization for fire predictions
3. We investigate the impacts of climate change on fuel load by the midcentury
4. We examine fire-induced changes in OC/BC by the midcentury

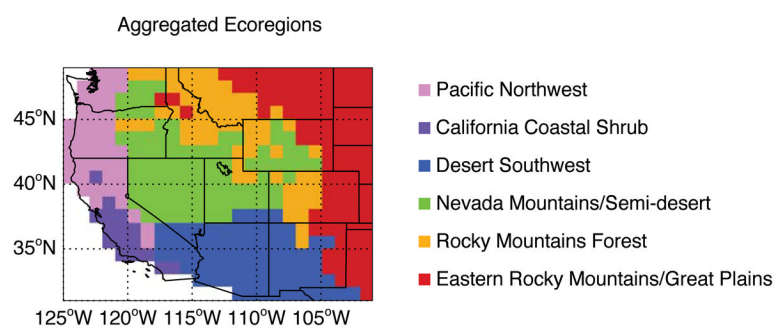
is the long-term average of the AB during fire season (May-October),  $\sigma$  is the corresponding standard deviation.



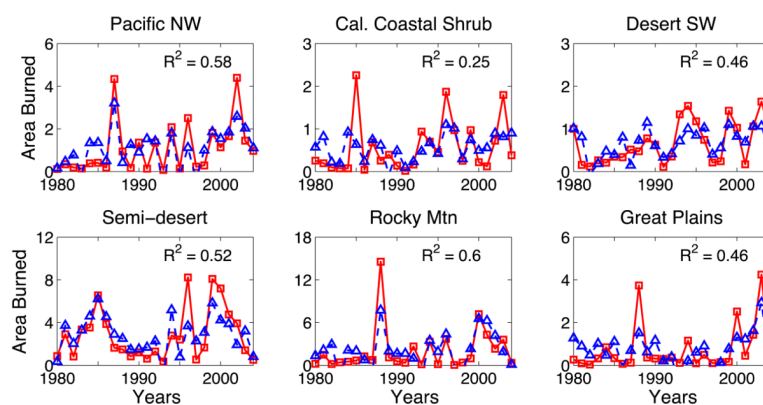
**Fig. 1.**

Illustration of the steps for the projection of future wildfire and surface aerosol concentrations of organic carbon (OC) and black carbon (BC) over the western United States in the present-day and at midcentury. The letters 'R' and 'P' denote the regression and parameterization methods. Site-based meteorological observations from the United States Historical Climatology Network (USHCN) and the Global Surface Summary of the Day (GSOD) are used for developing regression fits. The North American Regional Reanalysis (NARR) is applied for parameterization on  $1^{\circ} \times 1^{\circ}$  grids.



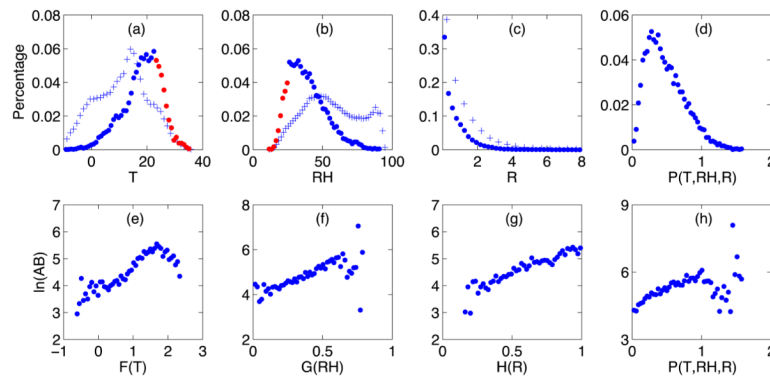


**Fig. 2.** Distribution of the six aggregated ecoregions over the western United States. Ecoregions are aggregated from 18 Bailey ecosystems following the method described in S2009.

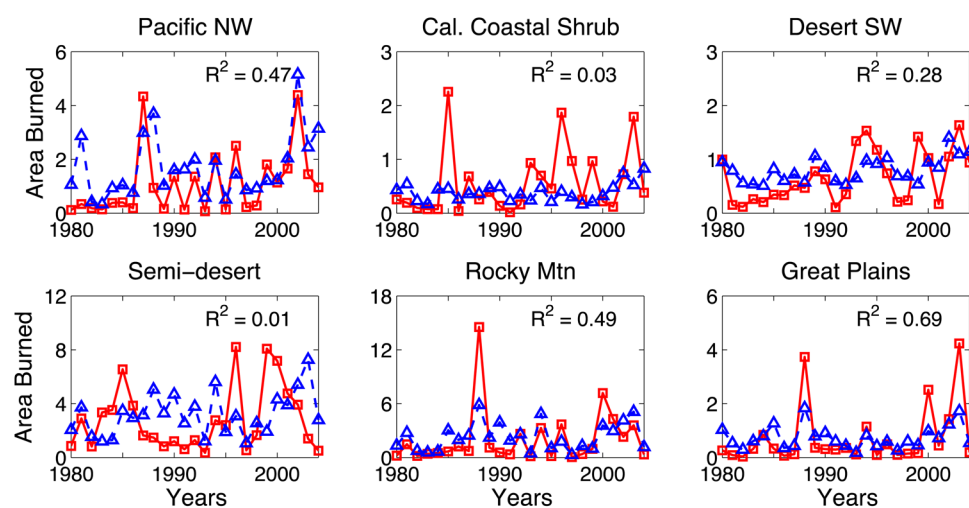


**Fig. 3.**

Observed (red solid lines) and predicted (blue dashed lines) area burned ( $10^5$  ha) for 1980–2004. The area burned is calculated using the regressions for the fire season (May–October) for each ecoregion. Site-based observations (USHCN and GSOD) are used in the prediction. The correlation  $R^2$  between observation and prediction is shown on each panel.

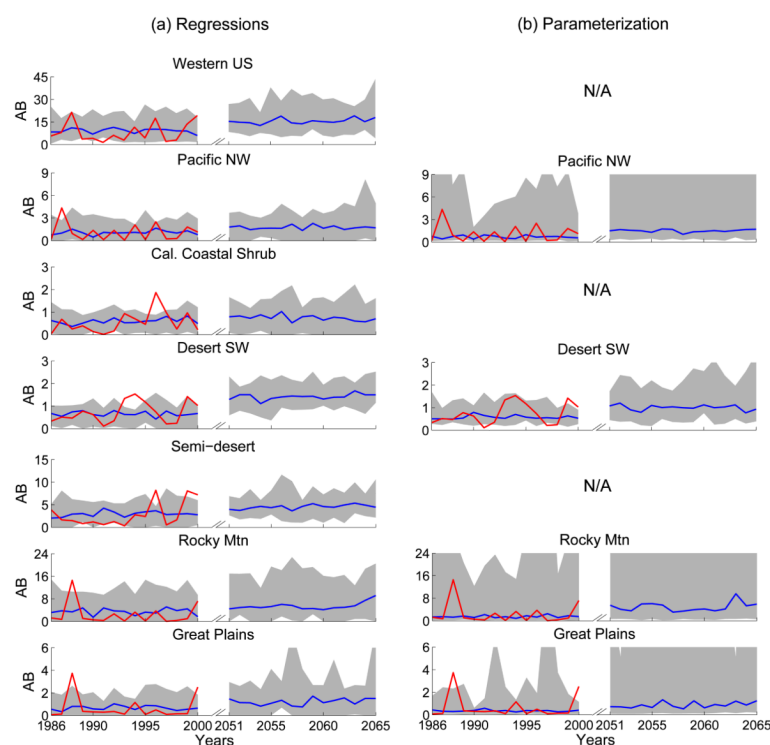
**Fig. 4.**

Probability distributions of monthly mean (a) surface air temperature ( $^{\circ}\text{C}$ ), (b) relative humidity (%), and (c) precipitation ( $\text{mm day}^{-1}$ ), for all grid cells in the western United States (plus signs) and for those cells that have observed monthly area burned larger than 10 ha (red and blue points) during 1980–2004. The red points in panels (a) and (b) indicate those fire cells with temperatures  $>23^{\circ}\text{C}$  and relative humidity  $<25\%$ , respectively. Panels (e)–(g) show the relationships between the logarithmic monthly area burned ( $\ln(\text{AB})$ , ha) and these functions: (e)  $F(T) = T/15$ , (f)  $G(\text{RH}) = (1 - \text{RH}/100)^2$ , and (g)  $H(R) = 1/(R+1)$ , where  $T$  is temperature,  $\text{RH}$  is relative humidity, and  $R$  is rainfall. Panel (d) shows the probability distribution for the function  $P(T, \text{RH}, R) = F(T) \times G(\text{RH}) \times H(R)$  over the same fire cells as in Panels (a)–(c). Panel (h) shows the relationship between  $\ln(\text{AB})$  and the function  $P(T, \text{RH}, R)$ . In all panels, the range for each meteorological variable or function has been divided into 50 equal-size bins, and the symbols represent the average of the cell values within each bin.



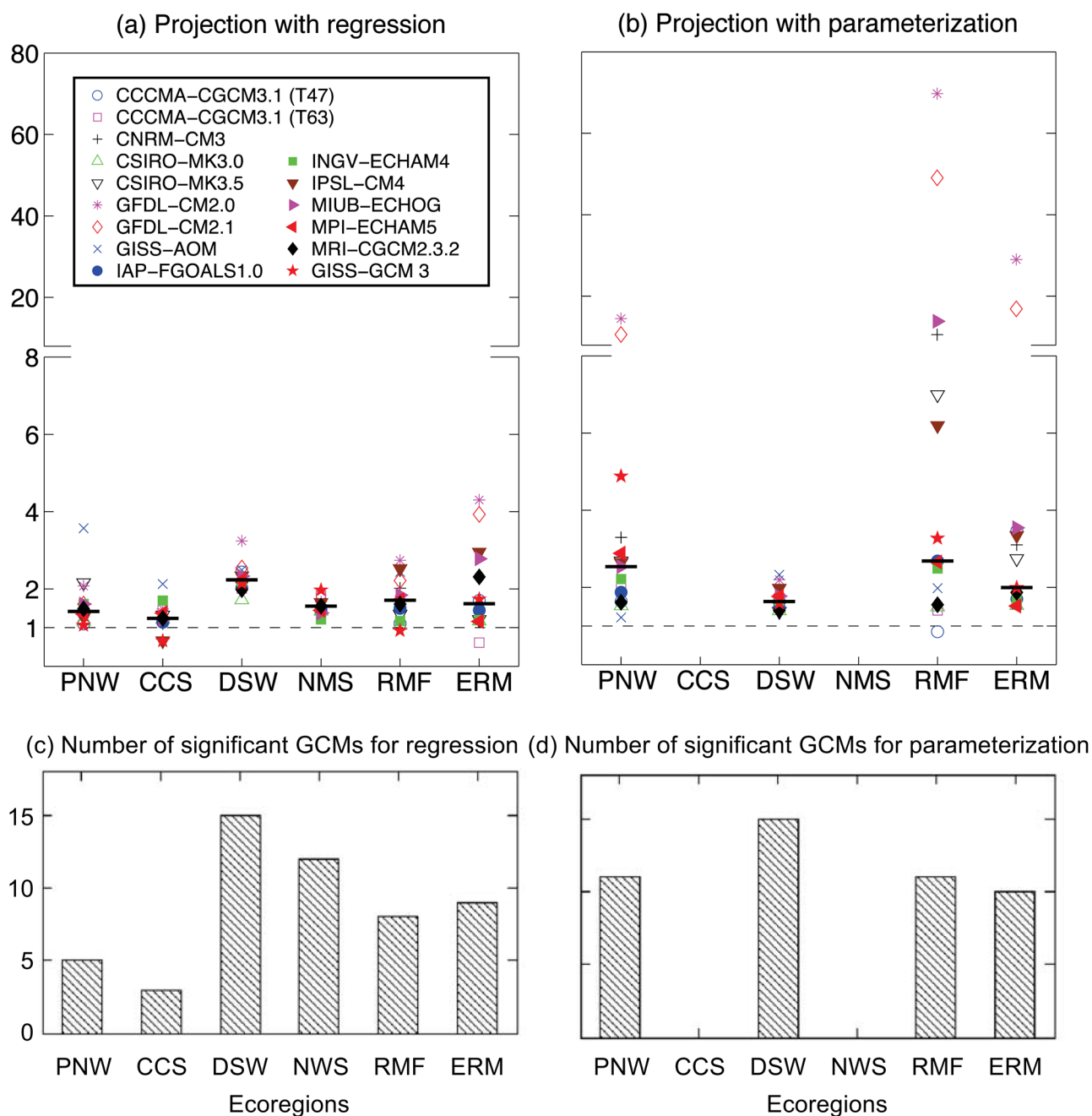
**Fig. 5.**

Time series of observed (red solid lines) and predicted (blue dashed lines) area burned ( $10^5$  ha) for 1980–2004 using the parameterization. The NARR gridded reanalyses are used in simulation.

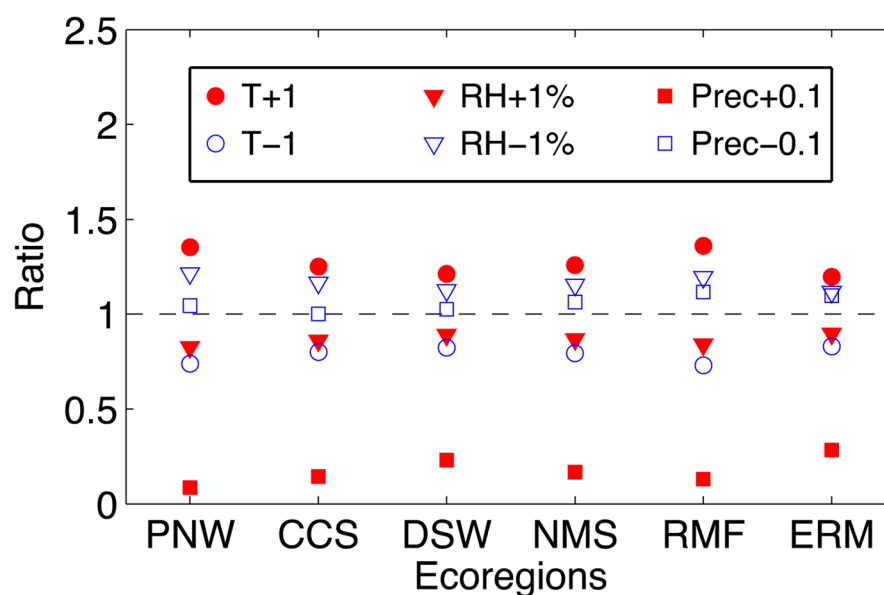


**Fig. 6.** Predicted annual mean area burned ( $10^5$  ha) for the present day (1986–2000) and at midcentury (2051–2065) from 15 GCMs using (a) regressions and (b) the parameterization. The red lines are the observed area burned for the western U.S. (top panel) or for each ecoregion (lower panels) during 1986–2000. The blue lines are the yearly median values of the area burned predicted by the 15 GCMs. The gray shading shows the spread of the predictions.

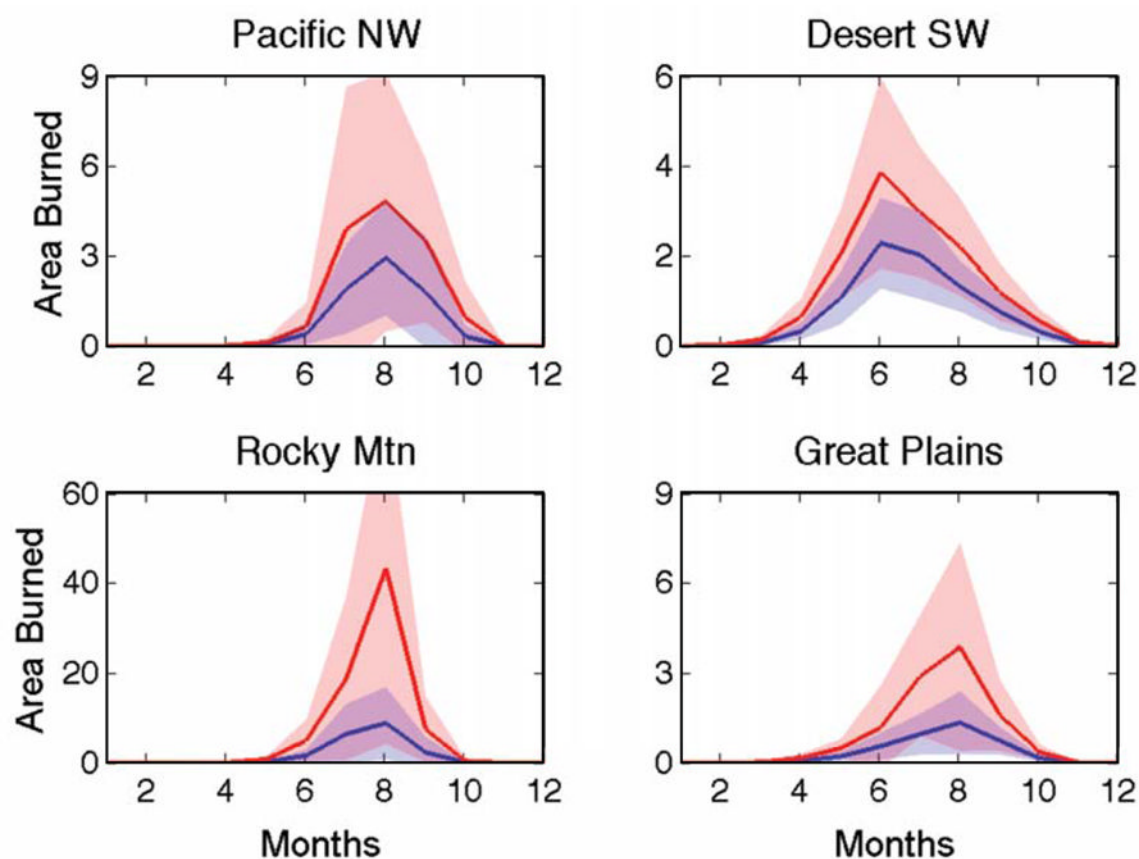


**Fig. 7.**

Predicted ratio of future area burned to present day area burned over six ecoregions with (a) regressions and (b) parameterization. The short bold lines are the median ratios. The number of GCMs that predict significant ( $p < 0.05$ ) increases in area burned is shown in (c) for regression and (d) for parameterization. The ecoregions are: PNW, Pacific Northwest; CCS, California Coastal Shrub; DSW, Desert Southwest; NMS, Nevada Mountains/Semi-desert; RMF, Rocky Mountains Forest; ERM, Eastern Rocky Mountains/Great Plains. Different symbols show results from different GCMs.

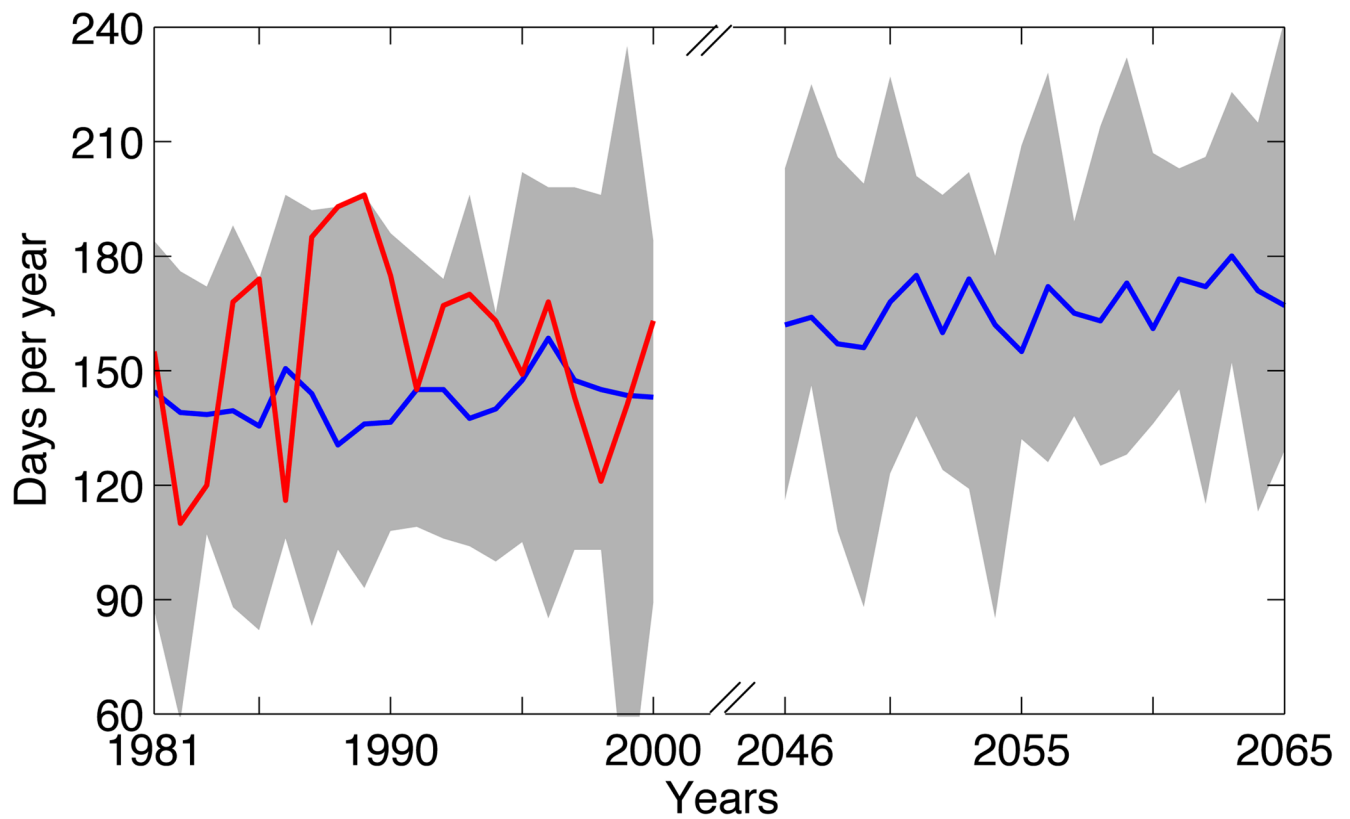


**Fig. 8.** Sensitivity of the parameterization to incremental changes in key meteorological variables. Values shown are the ratios of annual mean area burned calculated with perturbed variables to the area burned calculated with NARR meteorology, averaged over 1980–2004. Perturbations are to daily NARR temperature ( $\pm 1^\circ\text{C}$ ), relative humidity ( $\pm 1\%$ ), and precipitation ( $\pm 0.1 \text{ mm day}^{-1}$ ). The ecoregions are: PNW, Pacific Northwest; CCS, California Coastal Shrub; DSW, Desert Southwest; NMS, Nevada Mountains/Semi-desert; RMF, Rocky Mountains Forest; ERM, Eastern Rocky Mountains/Great Plains.



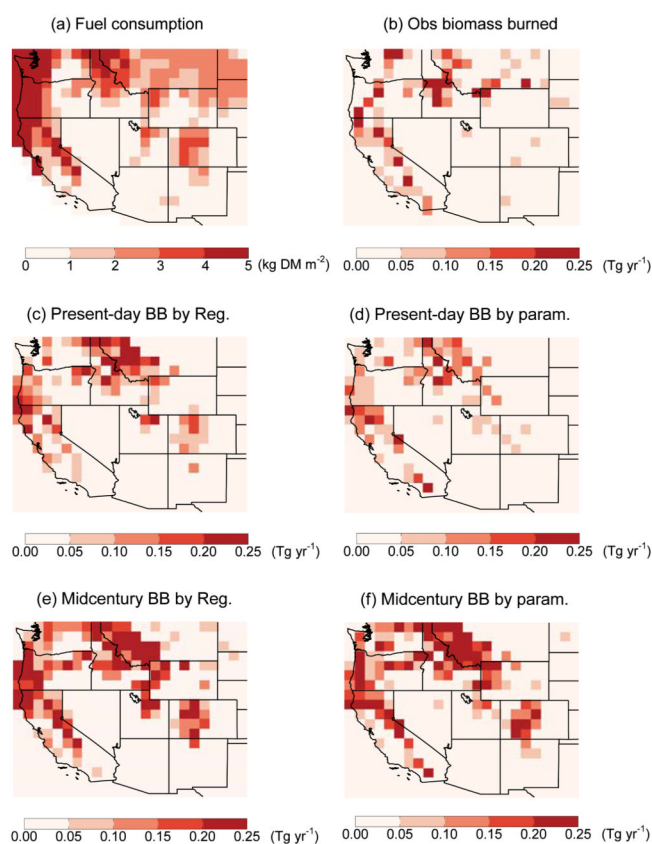
**Fig. 9.**

Predicted median area burned ( $10^4$  ha) each month for the present day (1981–2000, blue lines) and at midcentury (2046–2065, red lines) with meteorological variables from 15 GCMs, using the parameterization. The pink and purple shadings show the median standard deviation.



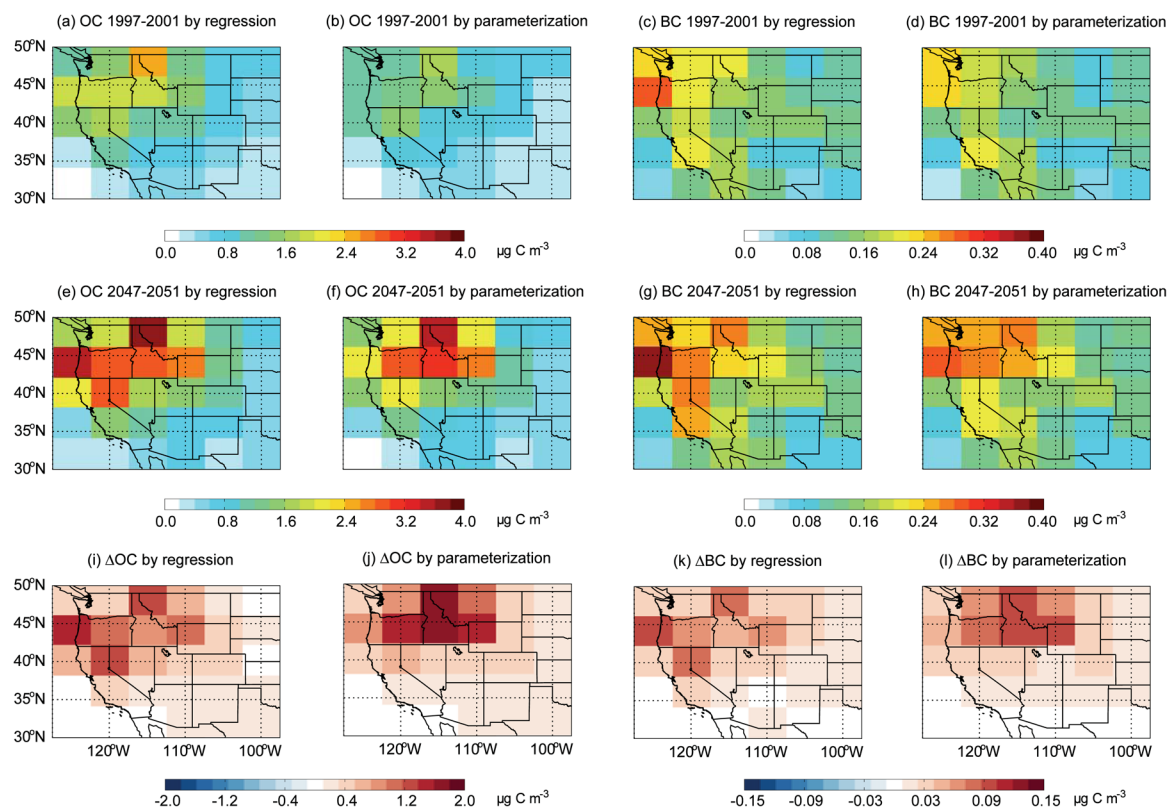
**Fig. 10.**

Predicted length of fire season for the present day (1981–2000) and midcentury (2046–2065) over western United States with 15 climate models. The fire season is defined as the period when the daily area burned is larger than 100 ha in at least one grid box ( $1^{\circ} \times 1^{\circ}$ ) over the four ecoregions in Fig. 9. The red lines are simulated results with NARR dataset for the present-day. The blue lines show the median length of fire season predicted with the 15 GCMs. The gray shading shows the spread of the predictions. Relative to the present-day medians, fire season lengthens by 23 days on average at the midcentury.

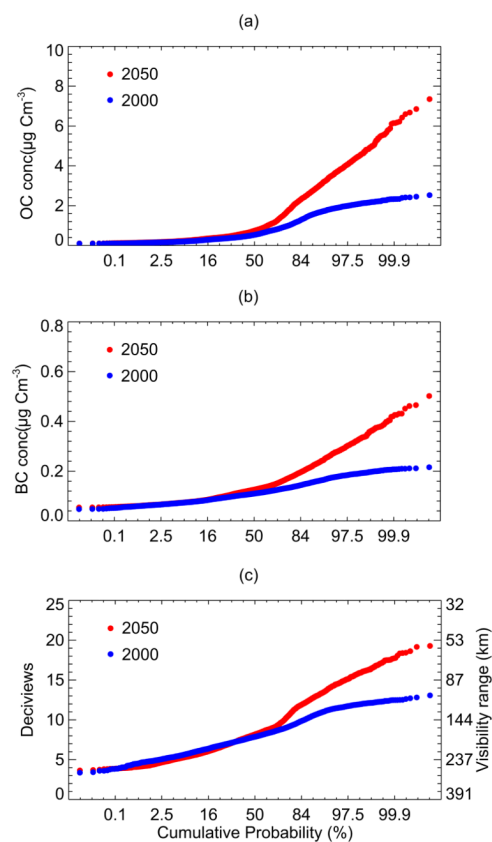
**Fig. 11.**

(a) Fuel consumption calculated from the Fuel Characteristic Classification System database. (b) Mean annual biomass burned during 1980–2004 based on observed area burned. Projected annual biomass burned for the present day (c, d) and midcentury (e, f) with regressions (c, e) or the parameterization (d, f). Results for regressions are during 1986–2000 for the present day and 2051–2065 for mid-century. For parameterization, the two periods are 1981–2000 and 2046–2065. We show different time periods for the two approaches since the regression model requires two years of antecedent meteorology.



**Fig. 12.**

Simulated June-August surface concentrations of (left 2 columns) OC and (right 2 columns) BC aerosol over the western United States during (a, b, c, d) 1997–2001 and (e, f, g, h) 2047–2051. Results are from the GEOS-Chem CTM using the median predicted fire emissions from the 15 GCMs with (1<sup>st</sup> and 3<sup>rd</sup> columns) regressions and (2<sup>nd</sup> and 4<sup>th</sup> columns) the parameterization. The bottom panel shows the change in simulated concentrations between the two time periods.



**Fig. 13.**

Cumulative probability distributions of daily mean (a) OC concentrations, (b) BC concentrations, and (c) haze index in deciviews in 32 Federal Class 1 areas in the Rocky Mountains Forest ecoregion in 2000 (blue points) and 2050 (red points). Each point represents average aerosol concentrations or haze index in one area on one day of the fire season (May-October). Emissions are calculated from the predictions using the parameterization.

**Table 1**Regression fits <sup>a</sup> for each aggregated ecoregion.

Ecoregion	Regression fits <sup>a</sup>	R <sup>2</sup>
Pacific Northwest	$3.4 \times 10^3 \text{ BUI} + 2.9 \times 10^4 \text{ RH.ANN}(-1) - 2.2 \times 10^6$	58%
California Coastal Shrub	$6.1 \times 10^3 \text{ RH.WIN}(-1) + 1.9 \times 10^2 \text{ DC}_{\text{max}} - 6.5 \times 10^5$	25%
Desert Southwest	$4.1 \times 10^4 \text{ T}_{\text{max}} \cdot \text{ANN} - 1.1 \times 10^4 \text{ T}_{\text{max}} \cdot \text{WIN}(-1) - 7.7 \times 10^5$	46%
Nevada Mtns/semidesert	$4.3 \times 10^4 \text{ RH.FS}(-1) + 8.6 \times 10^4 \text{ T}_{\text{max}} \cdot \text{SUM} + 3.8 \times 10^5 \text{ Prec.SUM}(-2) - 4.2 \times 10^6$	52%
Rocky Mtn Forest	$4.5 \times 10^3 \text{ DMC}_{\text{max}} - 7.4 \times 10^5$	60%
Eastern Rocky Mtns/Great Plains	$1.0 \times 10^3 \text{ DMC}_{\text{max}} - 2.0 \times 10^4 \text{ RH.ANN} + 1.0 \times 10^6$	46%

<sup>a</sup> A (-1) or (-2) after a predictor indicates that the meteorological field is 1 or 2 years earlier than current area burned. Variables are RH (relative humidity), T<sub>max</sub> (maximum temperature), Prec (precipitation), and fire indexes from Canadian Fire Weather Index system, such as Duff Moisture Code (DMC) and Build-up Index (BUI). Meteorological fields are averaged for winter (WIN, DJF), spring (SPR, MAY), summer (SUM, JJA), fire season (FS, MJJASO), and the whole year (ANN). The order of the terms indicates their contributions to the R<sup>2</sup> in the regression.

**Table 2**

List of models <sup>a</sup> whose outputs are used in future fire projection.

Model name	Resolution	Country
CCCMA-CGCM3.1 (T47)	3.75° × 3.75°	Canada
CCCMA-CGCM3.1 (T63)	2.8125° × 2.8125°	Canada
CNRM-CM3	2.8125° × 2.8125°	France
CSIRO-MK3.0	1.875° × 1.875°	Australia
CSIRO-MK3.5	1.875° × 1.875°	Australia
GFDL-CM2.0	2.5° × 2.0°	USA
GFDL-CM2.1	2.5° × 2.0°	USA
GISS-AOM	4.0° × 3.0°	USA
IAP-FGOALS1.0	2.8125° × 3.0°	China
INGV-ECHAM4	1.125° × 1.125°	Italy
IPSL-CM4	3.75° × 2.5°	France
MIUB-ECHOG	3.75° × 3.75°	Germany
MPI-ECHAM5	1.875° × 1.875°	Germany
MRI-CGCM2.3.2	2.8125° × 2.8125°	Japan
GISS-GCM3	5.0° × 4.0°	USA

<sup>a</sup>The meteorological fields for the first 14 models are from the CMIP3 archive (<http://www-pcmdi.llnl.gov/>). Other models in the archive did not provide required daily variables.

Table 3

Observations and ensemble projections of area burned ( $10^4$  ha) with regressions.

Ecoregions	Observed <sup>a</sup> (1986–2000)	Present Day Regression <sup>b</sup> (1986–2000)	Future Regression <sup>b</sup> (2051–2065)	Ratio <sup>c</sup> (Future/Present)	# of models <sup>d</sup> (p<0.05)	# of models <sup>e</sup> (M $\pm$ 20%)
Pacific Northwest	11.2 $\pm$ 12.0	11.3 $\pm$ 8.7	18.3 $\pm$ 10.2	1.42	5	10
California Coastal Shrub	5.4 $\pm$ 5.0	6.0 $\pm$ 3.2	7.0 $\pm$ 3.4	1.24	3	10
Desert Southwest	7.3 $\pm$ 4.7	6.5 $\pm$ 2.8	14.6 $\pm$ 3.2	2.24	15	13
Nevada Mnts/Semi-desert	28.1 $\pm$ 27.6	30.1 $\pm$ 18.5	48.9 $\pm$ 15.0	1.56	12	13
Rocky Mnts Forest	24.7 $\pm$ 38.6	35.9 $\pm$ 31.5	61.7 $\pm$ 33.0	1.71	8	6
Eastern Rocky Mnts/Great Plains	6.7 $\pm$ 10.6	6.9 $\pm$ 6.0	13.0 $\pm$ 6.5	1.62	9	5
Western U.S. total	83.3 $\pm$ 66.6	96.5 $\pm$ 50.9	155.3 $\pm$ 51.1	1.60	11	9

<sup>a</sup> AB = area burned. Results in each ecoregion are shown as  $\overline{AB} \pm \sigma \cdot \overline{AB}$

<sup>b</sup> Result in each ecoregion are the median values of  $\overline{AB}$  and  $\sigma$  predicted using the climate data from 15 GCMs for the A1B scenario.

<sup>c</sup> Results in each ecoregion are the median value of 15 ratios of AB between midcentury (A1B) and present-day calculated with the GCM output.

<sup>d</sup> Number out of 15 models who predict a significant (p<0.05) increase in AB in each ecoregion determined by a Student t-test.

<sup>e</sup> Number out of 15 models who predict a ratio within the range of  $\pm 20\%$  of the median ratio.

**Table 4**  
Observations and ensemble projections of area burned (10<sup>4</sup> ha) with the parameterization <sup>a</sup>.

Ecoregions	Observed (1981–2000)	Present Day Param. (1981–2000)	Future Param. (2046–2065)	Ratio (Future/Present)	# of models (p<0.05)	# of models (M±20%)
Pacific Northwest	9.1 ± 10.9	7.7 ± 5.0	19.8 ± 11.7	2.54	11	6
Desert Southwest	6.0 ± 4.7	6.2 ± 1.8	11.5 ± 3.3	1.63	15	12
Rocky Mnts Forest	20.2 ± 34.2	21.8 ± 15.8	98.8 ± 105.0	2.69	11	3
Eastern Rocky Mnts/Great Plains	5.8 ± 9.3	4.4 ± 2.1	14.0 ± 6.8	2.00	10	4

<sup>a</sup>Results are shown with the same format as in Table 3, but for a different present-day time period.



**Table 5**  
Changes of meteorological variables and their contributions to the predicted changes in area burned calculated by regressions.

Ecoregions	Simulated Median Mean		# of models (p<0.05) <sup>a</sup>	Changes in Reg. terms <sup>b</sup>	Percent Contribution <sup>c</sup>
	1986–2000	2051–2065			
Pacific Northwest					
BUI	81.3	105.0	10	$6.5 \times 10^4$	64
RH,ANN(-1) (%)	69.2	68.8	4	$-3.8 \times 10^4$	36
California Coastal Shrub					
RH,WIN(-1) (%)	68.5	67.8	2	$-2.2 \times 10^4$	37
DC <sub>max</sub>	1575.6	1654.4	10	$3.7 \times 10^4$	63
Desert Southwest					
T <sub>max</sub> -ANN (°C)	23.8	26.2	15	$9.3 \times 10^4$	89
T <sub>max</sub> -WIN(-1) (°C)	13.6	15.6	14	$-1.2 \times 10^4$	11
Nevada Mountains/Semi-desert					
RH,FS(-1) (%)	41.5	39.7	6	$-1.5 \times 10^5$	25
T <sub>max</sub> -SUM (°C)	29.1	31.7	15	$4.0 \times 10^5$	64
Prec.SUM(-2) (mm day <sup>-1</sup> )	0.62	0.57	4	$-0.7 \times 10^5$	11
Rocky Mountains Forest					
DMC <sub>max</sub>	238.9	298.7	8	$2.7 \times 10^5$	100
Eastern Rocky Mtns/Great Plains					
DMC <sub>max</sub>	118.3	155.3	7	$2.6 \times 10^4$	59
RH,ANN (%)	54.6	53.9	5	$1.8 \times 10^4$	41

<sup>a</sup>Number of models out of the 15 which predict significant (p<0.05) changes in meteorological variables in each ecoregion as determined by a Student t-test. If the median value of the change is positive, only those predicting a significant increase are counted and vice versa.

<sup>b</sup>Results are calculated as the changes in variables multiplied by the regression coefficients for the median models. A median model is defined as a model that predicts median ratios of the area burned in a specific ecoregion as shown in Table 3.

<sup>c</sup>Percent contributions of the absolute changes in individual regression terms to their sum for the median models.

Table 6

Annual mean dry biomass consumption from wildfire over western U.S.

Ecoregions	Annual Mean Biomass Consumption (Tg)					
	Observed (1980–2004)	Regression <sup>a</sup>			Parameterization <sup>a</sup>	
		1986–2000	2051–2065	1981–2000	1981–2000	2046–2065
Pacific Northwest	3.94	4.72	8.48	3.21		6.15
California Costal Shrub	0.64	0.75	0.87	N/A		N/A
Desert Southwest	0.34	0.25	0.55	0.25		0.44
Nevada Mountains/Semi-desert	1.22	1.72	2.47	N/A		N/A
Rocky Mountains Forest	5.05	8.63	15.6	5.09		18.0
Eastern Rocky Mountains/Great Plains	1.78	0.89	2.16	0.66		1.66
Western U.S. Total	13.0	17.0	30.1	12.1 <sup>b</sup>		29.1 <sup>b</sup>

<sup>a</sup>The years for regression and parameterization are different because the projected area burned with two methods have different time spans.

<sup>b</sup>Present-day (1981–2000) and midcentury (2046–2065) biomass burned over California Costal Shrub and Nevada Mountains/Semi-desert are calculated by using the observed monthly area burned from 1999–2004 in these two ecoregions.

**Table 7**List of numerical simulations with the GEOS-Chem CTM<sup>a</sup>

Simulations	Period	Fire emissions	
		Fire Model	Time scale <sup>b</sup>
PD.Reg	1997–2001	Regression	Monthly
PD.Param	1997–2001	Parameterization	Monthly
PD.Param.D	1997–2001	Parameterization	Daily
A1B.Reg	2047–2051	Regression	Monthly
A1B.Param	2047–2051	Parameterization	Monthly
A1B.Param.D	2047–2051	Parameterization	Daily

<sup>a</sup> All the simulations are carried out with the CTM driven by meteorological data archived from the GISS GCM3

<sup>b</sup> This is the temporal resolution of fire emissions (either monthly or daily) used in the simulations.